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A NEW FORM OF PUBLIC DISCLOSURE AND A CATALYST FOR CITIZEN ENGAGEMENT?

Emiliano Huet-Vaughn
Nicholas Muller
Yen-Chia Hsu

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Livestreaming Pollution: A New Form of Public Disclosure and a Catalyst for Citizen Engagement?
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ABSTRACT

Most environmental policy assumes the form of standards and enforcement. Scarce public budgets motivate the use of disclosure laws. This study explores a new form of pollution disclosure: real-time visual evidence of emissions provided on a free, public website. The paper tests whether the disclosure of visual evidence of emissions affects the nature and frequency of phone calls to the local air quality regulator. First, we test whether the presence of the camera affects the frequency of calls to the local air quality regulator about the facility monitored by the camera. Second, we test the relationship between the camera being active and the number of complaints about facilities other than the plant recorded by the camera. Our empirical results suggest that the camera did not affect the frequency of calls to the regulator about the monitored facility. However, the count of complaints pertaining to another prominent industrial polluter in the area, steel manufacturing plants, is positively associated with the camera being active. We propose two behavioral reasons for this finding: the prior knowledge hypothesis and affect heuristics. This study argues that visual evidence is a feasible approach to environmental oversight even during periods with diminished regulatory capacity.

Emiliano Huet-Vaughn
University of California – Los Angeles
4284 School of Public Affairs
Los Angeles, California 90095
ehuetvaughn@ucla.edu

Yen-Chia Hsu
Robotics Institute
Carnegie Mellon University
5000 Forbes Avenue
Pittsburgh, PA 15213
yenchiah@andrew.cmu.edu

Nicholas Muller
Department of Engineering, and Public Policy
Tepper School of Business
Carnegie Mellon University
Posner 254C
5000 Forbes Avenue
Pittsburgh, PA 15213
and NBER
nicholas.muller74@gmail.com

Introduction.

Environmental policy in developed economies depends on effective monitoring and enforcement. Monitoring ambient levels of pollution and direct mensuration of emissions are expensive and therefore incomplete.¹ A vast swath of the economy remains unmonitored. Enforcement, which is often contentious, requires the deployment of scarce public resources and is, therefore, imperfect. Recent developments in the federal political environment suggest even more limited federal enforcement (New York Times, 2017).

The dearth of pollution monitors and incomplete enforcement motivates public disclosure as a means to affect change among polluting firms (Tietenberg, 1998). Direct information on pollution provided to concerned citizens may enhance public pressure on regulatory agencies to enforce existing rules and laws. Traditional approaches to public disclosure in the environmental realm manifest as legal requirements for firms to reveal, or list, their emissions (Graham, 2002). Prominent examples of disclosure laws include the Toxic Release Inventory (TRI), and the Greenhouse Gas Reporting Program (GHGRP), (40 CFR Part 98).

¹ The current network of ambient air pollution monitors in the United States (see <https://www.epa.gov/outdoor-air-quality-data>) is sparsely distributed. Monitoring sites are typically chosen to maximize the likelihood of detecting a violation. Hence, the monitors are clustered in densely populated locations. Further, the Continuous Emissions Monitoring System (CEMS), an example of a system of emissions monitors, only tracks discharges of nitrogen oxides (NO_x) and sulfur dioxide (SO₂) at certain point sources.

This study focuses on a new means to provide the public with information on emissions and ambient pollution: visual evidence of emissions captured by camera, and broadcast on a free, public website. Essentially, livestreaming images of emissions is a means of disclosing risk to the public. In contrast to perhaps more conventional means (such as periodically reporting pollution levels or announcing risks to certain population groups) broadcasting live imagery of emissions provides a contemporaneous view of the threat from pollution in a more graphic, tangible, and visceral manner. Previous research suggests that people react to risk messaging in two general ways: systematic, or cognitive, responses and emotional responses (Lowenstein and Mather, 1990; Lowenstein et al., 2001; Slovic et al., 2002; Dillard and Anderson, 2004; Hastings, Stead, and Webb, 2004). Providing quantitative risk estimates - like direct measurements of pollution data - clearly speaks to cognitive message processing. The camera targets emotional reactions - anger or fear - that may trigger acute responses such as picking up the phone and calling the local air quality regulator (Averbeck, Jones, and Robertson, 2011).

Undergirding our analysis is the argument that visual disclosure may present a new set of tools for regulators, environmental advocates, and concerned communities to affect behavior of firms that manage to circumvent traditional regulatory frameworks.

Recognizing the potential for visual evidence as an important complement to customary monitoring and enforcement, Giles (2013) notes that:

“[W]e are not far from the day when the public will have access to pollution monitoring tools. Communities with monitoring data will encourage better

performance by industries they host...These changes, driven by new technologies, will encourage more direct industry and community engagement, and reduce the need for government action.”

In addition to its comprehensivity, particularly attractive is the cost-effectiveness of this approach. Any demonstrable reduction in pollution attained through this approach would come at very low opportunity cost, relative to more traditional approaches that require significant allocations of physical and financial capital as well as labor resources.²

1.1 Empirical Context

The analysis focuses on a now shuttered coke plant located just west of Pittsburgh, Pennsylvania on Neville Island in the Ohio River: the Shenango Coke Works, or the “mill”. Coke is purified coal; it is a nearly pure carbon substance produced by baking coal to remove impurities. This site has repeatedly violated air pollution standards.³

The mill has received considerable attention in the local media and in nearby communities (Pittsburgh Post-Gazette, 2015a).⁴ A robotic camera was installed to track

² Over the past twenty years, there were over 19,000 inspections performed by the United States Environmental Protection Agency (USEPA), per year. Of these, about 1,000 to 3,000 were related to the Clean Air Act (CAA), (Shimshak, 2014). The USEPA’s Office of Enforcement and Compliance Assurance has an annual budget in the range of \$600 million (Shimshak, 2014).

³ Recent examples of such violations include: Allegheny County (home to the plant) fined DTE Energy (the corporate owner of the plant) in late 2013. In the spring of 2014, the operator of the plant reached a settlement with the Allegheny County Health Department. The agreement contains specific changes to operations designed to reduce emissions, investments in pollution control technology, and fines related to earlier violations (Pittsburgh Tribune Review, 2014). Previously, in 2012, DTE reached a consent decree involving a \$1.75 million penalty (Pittsburgh Post-Gazette, 2015b).

⁴ In particular, violations occurred in: 1980, 1993, 2000, 2005, 2012, and 2014.

visible emissions from the plant (Shenango Channel, 2015; Pittsburgh Post-Gazette, 2015b). The Shenango Channel first went live on November 15th, 2014, providing visual smoke imagery online for most of the next three weeks until stopping the broadcast on December 5th, 2014. Then, again, on January 22nd, 2015 the livestream began anew and continued to broadcast for the duration of 2015. The imagery captured by the camera is posted on a website devoted to providing visual evidence of emissions from the mill (Shenango Channel, 2015). The device also generates quantitative estimates of the opacity of emissions from the facility every five minutes. Emissions density is quantified through pixel counts. Thus, in addition to the raw visual imagery posted online, the camera provides a numerically based approach to real-time measurement of emissions that is both more reliable than ad hoc observations made by members of affected communities, and distinct from mensuration performed by regulators.

The empirical analysis begins with a test of whether there is an association between the quantitative smoke readings produced by the camera's algorithm and actual pollution observations gathered at a nearby (ambient) pollution monitoring site. This first set of tests examine the internal validity of the visual data. That is: do the smoke images reflect actual conditions in nearby communities? As such, are the data then useful as an additional means of environmental monitoring?

The second set of empirical exercises tests whether the visual evidence provided on the internet affects public engagement with the local air quality regulator. By extension, we argue that the degree of engagement is a reflection of how the citizenry perceives risk

from air pollution. We begin by testing whether the number of air pollution-related calls to the local air quality regulator (the Allegheny County Health Department, or ACHD) is associated with the activation and continued operation of the camera and the visual data published online.

We then conduct two additional tests that emanate from the literature on behavioral responses to risk messaging in the following way. A key determinant of how individuals react to information is prior knowledge about the subject matter (Averbeck, Jones, and Robertson, 2011). If an intervention conveys threat information that for some subjects is utterly new, while for others it is known, it is intuitive that the emotional effect on the latter groups will be less than the former. Without an existing knowledge base, the uninformed go with their gut instinct.

In order to parse calls according to prior knowledge held by the caller, we subdivide calls to the ACHD in two ways. First, we test whether the number of calls to the ACHD that refer to the Shenango mill is associated with the presence of the camera and the visual data published online. Calls targeting Shenango overwhelmingly originate from the zip code that contains the mill, both before and after the deployment of the camera (see ACHD, 2016). Hence, prior knowledge is high among these callers.

Next, we limit the sample to calls that single out steel mills. (While steel-related calls are the second most common industrial category of calls in the ACHD air pollution complaint data, there are no steel manufacturing facilities in the zip code that contains Shenango.) Thus, these callers are much less likely to experience the pollution from the

mill regularly. The prior knowledge hypothesis suggests that onset of the camera imagery is likely to generate a stronger response among these communities if only because it is a true shock to their existing knowledge.

Also relevant to our empirical design is a phenomenon known as the affect heuristic: an emotionally based short cut to decision-making (Finucane, et al., 2000; Slovic et al., 2002; Kahneman and Frederick, 2002; Shiller, 2017), which the literature emphasizes is salient to risk assessments (Keller, Siegrist, and Gutscher, 2006). One manifestation of affect heuristics is that people experiencing strong emotional responses to a stimulus may apply their emotions to other circumstances (Shiller, 2017). For example, a person not in the direct vicinity of the mill accesses the livestreaming website and responds emotionally to what they see. When they observe emissions in their own neighborhood, they are more likely to call the ACHD because of their heightened emotional state. Both the affect heuristic and the prior knowledge hypothesis provide a behavioral basis for the camera influencing public perception about the risk from air pollution more broadly than just centered on the mill. Such responses bolster the case that visual monitoring and real-time disclosure of imagery may serve as a broad-based strategy to raise awareness and, subsequently, boost citizen engagement with local environmental enforcement authorities.

This paper relates to several aspects of the literature in economics. First, it builds on research exploring disclosure laws (Konar and Cohen, 1997; Tietenberg, 1998; Afsah, Blackman, Ratunanda, 2000; Cohen and Santhakumar, 2007; Garcia, Sterner, and Afsah,

2007; Blackman, 2010; Huang and Kung, 2010). For a discussion of emission reductions from such programs see (Konar and Cohen 1997; Foulon et al., 2002; Hahn et al., 2003). Since we study the use of the camera as a means to monitor emissions from a point source of pollution, the paper also is associated with the literature on enforcement and monitoring. Shimshak (2014) provides an overview of enforcement and monitoring of environmental laws. Finally, the notion that the public may have concerns over emissions and concentrations of ambient pollution stems, in part, from a literature that reports and association between exposure and adverse health impacts (see for example, Krewski et al., 2009; LePeule et al., 2013). Prior research in the policy literature discusses advanced pollution monitoring techniques that pertain to the present study, (Giles, 2013). The methods used to collect visual evidence are discussed briefly herein and pertain to a set of techniques discussed in Hsu et al., (2017).

1.2 Preview of Results

The empirical analysis finds statistical evidence that the visual smoke data are associated with PM_{2.5} levels at the nearby monitoring station operated by the ACHD as part of USEPA's network. The statistical association is strongest when controlling for wind direction: a one-hour lagged value of the pixelated smoke data interacted with wind direction is significantly associated with PM_{2.5} levels at the ambient pollution monitoring station. The use of a lagged measure of pollution reflects the difference between the real-time measurement of smoke imagery and the time it takes for

emissions to reach the monitoring station, which depends on both wind speed and direction.

We detect evidence of an association between the daily count of all calls pertaining to air pollution in the ACHD's database and the presence of the camera. In the most parsimonious specification, the camera being on is associated with an increase of one call every two days ($p < 0.01$). In our preferred specification, we find evidence that interactions between the active camera indicator and day of the week fixed effects are significantly associated with call counts. For example, the interaction with the Monday fixed effect suggests an increase of 0.3 calls per day ($p < 0.05$), relative to the count of calls on weekend days, above and beyond this Monday-weekend difference prior to the onset of the camera. The interaction between the camera control and other weekdays indicates that the camera produces an increase twice as large ($p < 0.01$), again, relative to weekend days, as compared to this day-of-week difference prior to the deployment of the camera.

The models that feature complaints about Shenango, the monitored facility, suggest that the camera had no effect on daily Shenango-related call counts. We control for the announcement (in December of 2015) of the closure of Shenango, and find that this event resulted in a permanent reduction of about two calls per day ($p < 0.01$).

Our final set of tests explores the association between the camera and complaints targeting steel manufacturing facilities. In our preferred specification, we detect evidence of an effect through the interaction of the active camera indicator with day-of-

week controls; the camera boosted the pre-camera Monday peak in calls and diminished the pre-camera tapering of calls in the rest of workweek. For example, the interaction term with the Monday control suggests that the camera was associated with an increase of one call per day, relative to weekend days, above and beyond this Monday-weekend difference prior to the onset of the camera ($p < 0.01$). In what is perhaps evidence of a persistent effect of the camera on citizen behavior, the Shenango closure announcement does not affect the call counts targeting steel mills.

The remainder of the paper is organized as follows. Section 2 describes the data, the approaches used to gather the visual smoke data and our econometric modeling. Section 3 explores the results while section 4 concludes.

2. Data and Method

This section is subdivided into three parts. The first describes data used in the analysis. The second subsection explores the approach to gathering and analyzing the visual smoke emissions data. The third subpart discusses the econometric techniques and model specifications.

2.1 Data

Hourly observations of ambient pollution are obtained from the USEPA Air Quality System (AQS) database. These data are provided on an hourly basis, which are then aggregated up to the day when used in conjunction with the daily call counts. The monitoring data includes fine particulate matter ($PM_{2.5}$), sulfur dioxide (SO_2), ozone

(O₃), and for the Avalon air quality monitor operated by the ACHD, hydrogen sulfide (H₂S). Data are gathered for 2014 through 2016. Hourly weather data, also aggregated up to the day when used with the daily call count data, are also assembled from a nearby weather station. These include wind speed, wind direction, ambient outdoor temperature, and atmospheric pressure. The livestream smoke resolution data comes from collaborative initiative between Carnegie Mellon University researchers and Pittsburgh community members (as described in greater detail in Section 2.2).

Table 1 summarizes these data. The average PM_{2.5} level across all monitors in Allegheny County over the time period under examination was 15.28 ug/m³. At the Avalon monitor, PM_{2.5} averages 12.6 ug/m³. The hourly maximum reading of SO₂ averaged 21.77 ppb. The hourly maximum O₃ level was 41.42 ppb. (We report maximums for SO₂ and O₃ because the National Ambient Air Quality Standards set by the Clean Air Act are defined in terms of maximum values.) H₂S averaged less than 1 ppb. The mean temperature at the monitor location is about 18°C. The site, which is located in the Ohio River Valley, is not characterized by high winds; average wind speed is just 3.8 miles per hour. Table 1 indicates that the mean wind direction is southwesterly (204°). However, figure A.1, which shows a histogram of wind direction, indicates that the distribution of wind direction is multi-modal. Winds most frequently blow from between 250° and 360°. Thus, prevailing winds are westerly and northwesterly. The smoke resolution data averages 72 pixels out of 4,000 pixels in a given frame. The maximum smoke reading is 2,666 pixels. There are numerous zeroes in the data

corresponding to hours in which there is no smoke detected: 2,625 out of 3,061 observations are zero.

Data on public air quality complaints to government regulators come from the ACHD Air Quality Program, which is responsible for “regulating air pollutants” as well as “enforcing federal pollution standards, and permitting industrial sources of air pollution” within Allegheny County, site of the Shenango Coke Mill (ACHD Air Quality Program, 2016). The complaints are made either by phone or online and recorded by ACHD employees with detailed information about the time and nature of the complaint, as well as a categorization of the offending party the caller is complaining about (when discernable) and zip code locations of both the caller and alleged offender coded when possible.

Of 2,314 total complaints made in 2014 and 2015, 944 clearly are marked as pertaining to the Shenango Coke Mill, representing nearly 41% of all air quality complaints, by far the largest cause of air quality complaints to ACHD in this time period. The next most common complaint source categorized by ACHD is steel manufacturing (115 complaints, or, 5% of all complaints). Many of the remaining complaints are not directly related to specific industrial sources of pollution either because the complaint is about general air quality (163 complaints, or, 7% of all complaints) or odors that are not source-specific (403 complaints, or, 17.6% of all complaints). In addition, some of the complaints deal with pollution typically caused by fellow citizens or seasonal allergens rather than industry: for example, open burning (161, or, 7% of all complaints), wood

smoke (162, or, 7% of all complaints), and, dust (103 complaints, or, 4.5% of all complaints).

Over the 730 days encompassed by the analysis, there were 3.5 calls per day pertaining to air pollution received by the ACHD. The maximum daily call volume was 31. There were 1.3 calls per day specifically related to the Shenango Mill. The maximum Shenango-related call count was 16. And, there were 0.1 calls per day focused on steel mills.

As documented in Figure 1, there is a clear day-of-week trend in the ACHD complaints. The top panel of Figure 1 shows all calls occurring before (bottom red line) and after (top blue line) the installation of the camera. With and without the camera, call volumes are highest on Mondays. The call frequency tapers off during the remainder of the week. This pattern is also evident for the calls about Shenango. Steel mill-related calls also spike on Monday and taper off during the rest of the week in the period without the camera, but, when the camera is on, the average call count rises at the end of the work week (Thursday and Friday actually are the highest average call count days). For all three groups of call types, we observe higher call volumes after the camera was installed for weekdays. The clear day of week effects and the apparent reduction in the degree to which calls taper off later in the workweek when the camera livestream is on motivate day of week fixed effects and day of week-camera interactions in the statistical analysis to follow.

2.2 Assessment of Visual Smoke Emissions.

2.2.1. Gathering Video Data

The smoke data are gathered through a collaboration with a local community in Pittsburgh, Pennsylvania to document images of fugitive emissions from the Shenango mill. Starting from November 2014, researchers at Carnegie Mellon University have helped the local community build a live camera monitoring system (Shenango Channel, 2015; Hsu et al., 2017) pointing at the coke oven where the fugitive emissions usually happen. The camera takes a picture every 5 seconds and gathers nearly 17,000 images for one day. The system processes the imagery gathered each day into a time-lapse video and visualizes the result by using a web-based large-scale time-lapse viewer, available in real time, which was developed previously (Sargent et al., 2010). The interactive viewer (see the top-left panel of Figure A.4 in the appendix; Figure A.4 presents what users see on the website) facilitates the exploration of high quality time-series images by panning and zooming for finding fugitive emissions.

2.2.2. Smoke Detection Algorithm

It is important to re-emphasize that the camera serves two purposes. First, it is used to broadcast real-time imagery of emissions on the web. Second, it is used to produce quantitative estimates of visible particulate emissions produced by the mill. To measure emissions, the system provides a thumbnail tool for generating and sharing animated smoke images. However, manually searching through each image to identify smoke emissions is prohibitively inefficient for computational purposes. As such, we have

implemented a computer vision tool that uses a baseline smoke detection algorithm for detecting industrial smoke emissions during the daytime and generating related animated images automatically. Figure A.2 in the appendix demonstrates various smoke emission images. It also shows steam, shadows, and the mixture of steam and smoke that may confound smoke images. The task is to detect frames from a static camera containing smoke, exclude the ones having steam and shadow, identify the starting and ending frames of emissions, and output animated images that include smoke used for quantification. The following subsections describe three main steps of the algorithm: change detection, texture segmentation, and region filtering. Figure A.3 in the appendix outlines the steps.

2.2.3. Change Detection

The purpose of change detection is to identify moving pixels that may contain smoke. Smoke is semi-transparent with various opacities and occludes parts of the background upon presence, which causes changes of high frequency signals and pixel intensity values across frames. To reduce the computational cost, we first scale the original image at time (t) down to one-fourth of the original size to obtain a down sampled image denoted (I_t) in Figure A.3. Next, we estimate the background image (B_t) by taking the median over the previous 60 images. Then we subtract the pixel intensity values in the estimated background image from the current image to get a residual image and threshold the residual image to obtain a binary mask (shown as M_{heq} in Figure A.3). We also filter high frequency signals in (I_t) and (B_t) and perform the same background

subtraction process to obtain another binary mask (M_{dog}). Finally we combine (M_{heq}) and (M_{dog}) into (M_{cd}) which indicates moving pixels.

2.2.4 Texture Segmentation

Texture segmentation clusters pixels into several candidate regions based on texture information. We first convolve the current image (I_t) with a filter bank (a set of 5-by-5 convolution masks) to obtain feature vectors. Each vector represents the corresponding pixel in a high dimensional space. Then we perform Principal Component Analysis that preserves 98% of the energy (eigenvalues) on the feature vectors to reduce dimensions. Finally we run a k-means++ algorithm which chooses better initialized values (seed points) to cluster feature vectors into textons. We use these textons to divide the current image (I_t) into various regions as shown in image (R_t).

2.2.5 Region Filtering

Region filtering iteratively evaluates each candidate region based on shape, color, size, and the amount of changes to determine if it matches the appearance and behavior of smoke. We first smooth the image (R_t) by discarding small regions, removing noise by using a median filter, and performing morphological closing. Next, we use the connected component algorithm to find all separated regions and remove the ones that are thin and narrow. Then we group nearby regions having white or black colors to reconstruct the shapes of objects. Since the color of smoke is usually grayish or bluish, we can remove regions having non-grayish and non-bluish colors. We also exclude regions having extremely light colors because steam is usually white. Then we compute the size of each region and ignore extremely large or small ones that may be noise and

shadow respectively. Furthermore, we eliminate regions having insufficient amount of moving pixels based on image (M_{cd}). Finally, we remove regions that may contain shadow by using a baseline shadow detection algorithm. (R_{filter}) shows the final result, and (M_t) indicates the union of smoke regions.

2.2.5 Visualization

The computer vision tool provides three visualization features: an interactive timeline for video seeking, an autonomous fast-forwarding feature for skipping uninteresting frames, and a visual summary of animated images that are likely to contain smoke for documentation. We first use the smoke detection algorithm to predict the number of smoke pixels in a video frame (see the top graph in Figure A.5). The x-axis and y-axis indicate the frame number and the sum of smoke pixels in a frame respectively. Next, we compute the peaks and the corresponding peak widths to obtain frame segments (see Figure A.5). Then we visualize the graph using an interactive timeline (see the bottom-left graph in Figure A.4), which gives indicators of emissions.

The top panel of figure A.5 shows (graphically) the pixel counts that are subsequently used in the econometric analysis. The time signature for the smoke readings enable joining to weather and pollution data.

2.3 Econometric Analysis

The first set of empirical analyses focus on the determinants of ambient pollution ($PM_{2.5}$, SO_2 , H_2S) at the Avalon monitor that is situated very near to the Shenango mill. Within this category of tests, the central hypothesis test is whether there is an

association between the visual smoke readings and ambient PM_{2.5}, SO₂, and H₂S at the Avalon monitor. In effect, we test the internal validity of the information captured by the camera with respect to more traditionally gathered data on ambient air pollution. Because the visual emissions data, by definition, are visible, we expect that the only plausible empirical relationship is between smoke and ambient PM_{2.5} since SO₂ and H₂S are gaseous. Additional controls in the models include the determinants of ambient pollution: wind speed and direction, a linear time trend, month, day, and hour of the day, temperature, pressure, as well as day of the week.

The second group of hypotheses explore whether there is a public response to the information provided by the digital camera. We test for an association between all calls to the ACHD, calls targeting the Shenango mill, and complaints that zero-in on steel mills.

2.3.1 Determinants of Ambient PM_{2.5}

Model (1) is the default specification and all covariates enter in linear and quadratic forms. For each pollutant, we estimate the following model.

$$P_t = \beta_0 + \beta_1 T_t + \beta_2 W_t + \beta_3 E_t + \beta_4 S_t + \varepsilon_t \quad (1)$$

where: P_t = ambient pollution at the Avalon monitor at time (t).

T_t = time controls: hour, day, month, and linear time trend at time (t).

W_t = weather controls: wind speed, direction, temperature, and pressure at time (t).

E_t = environmental pollutants other than the dependent variable at time (t).

S_t = smoke readings from camera at time (t).

ε_t = idiosyncratic error term.

In the context of model (1), the primary hypothesis test focuses on β_4 . That is:

$$\begin{aligned} H_0: \beta_4 &= 0 \\ H_A: \beta_4 &\neq 0 \end{aligned}$$

Model (2) includes interaction terms between wind direction and the smoke readings.

Let $S_t^W = (S_t \times \text{Wind Direction}_t)$. Then model (2) is given by:

$$P_t = \beta_0 + \beta_1 T_t + \beta_2 W_t + \beta_3 E_t + \beta_4 S_t + \beta_5 S_t^W + \varepsilon_t \quad (2)$$

Model (3) recognizes that, given the low observed wind speeds, it may take time between when smoke is detected visually (by the camera) and when an effect of such smoke registers at the air quality monitoring station across the river. To incorporate the potential delay between smoke emission and ambient readings, model (3) includes a one-hour lagged measure of smoke releases.

$$P_t = \beta_0 + \beta_1 T_t + \beta_2 W_t + \beta_3 E_t + \beta_4 S_t + \beta_5 S_t^W + \beta_6 S_{t-1} + \beta_7 S_{t-1}^W + \varepsilon_t \quad (3)$$

Model (4) tests whether extreme episodic emissions from the Shenango Coke Works had a measureable effect on $PM_{2.5}$ readings. In particular, this specification includes a dummy variable which assumes the value of unity (zero otherwise) for all days between May 26th through June 15th, 2015. These dates correspond to a series of power outages at the facility that also resulted in fires at the plant. The incident dummy variable is also interacted with wind direction. Let $I_t^W = (I_t \times \text{Wind Direction}_t)$. Thus, model (4) is given by:

$$P_t = \beta_0 + \beta_1 T_t + \beta_2 W_t + \beta_3 E_t + \beta_4 S_t + \beta_5 S_t^W + \beta_6 S_{t-1} + \beta_7 S_{t-1}^W + \beta_8 I_t + \beta_9 I_t^W + \varepsilon_t \quad (4)$$

2.3.2. Public Response to the Camera.

The second dimension of the empirical analysis tests whether the installation of the camera surveillance system trained on the plant affects the frequency and nature of complaints made to ACHD. We employ three different dependent variables. The first is the daily count of all calls about air pollution to the ACHD. The second dependent variable is the number of complaints that clearly refer to the Shenango mill. The third outcome variable is the daily count of calls that single out steel mills. In models (5), (6), and (7) shown below, each of these three dependent variables are employed. These models are estimated using OLS, Poisson, and negative binomial estimators; our default results feature the negative binomial estimator as it is well known that OLS is inappropriate in count data contexts. The dependent variables show evidence of over dispersion, which detracts from the viability of the Poisson estimator.

Model (5), our most parsimonious specification, regresses daily ACHD calls on two indicator variables: one for the initial days during which the camera and livestreaming footage was made public⁵, and one for all days during which the camera was actively streaming images to the internet.

$$C_t = \alpha_0 + \alpha_1 O_t + \alpha_2 A_t + v_t \quad (5)$$

⁵ The camera was initially turned off and on over a period of days, and didn't remain permanently on until January of 2015, so, there are two weeks we treat as onset weeks.

Where: C_t = count of complaints/posts in week (t).

O_t = indicator for days of camera and data onset.

A_t = indicator for days during which camera and data are active.

v_t = stochastic error term.

To this specification we subsequently add season fixed effects, a cubic time trend, and a suite of pollution controls from the monitors in Allegheny County. We also add covariates for weather conditions: temperature, wind speed, wind direction, and pressure. Model (6) also controls for the days during which local newspapers ran articles on the camera and the accompanying website, as well as the day on which the ACHD announced that it was in talks with the USEPA about executing a new enforcement action against the mill; this day also coincided with a community meeting held by a local environmental activist organization. This specification also controls for the announcement of the closure of the Shenango facility, which happened in December of 2015.

$$C_t = \alpha_0 + \alpha_1 O_t + \alpha_2 A_t + \alpha_3 P_t + \alpha_4 W_t + \alpha_5 T_t + \alpha_6 N_t + \alpha_7 E_t + v_t \quad (6)$$

Model (7) allows for interactions between the active camera indicator variable and pollution readings. Model (8) extends model (7) to include day-of-week fixed effects as well as interactions between the camera indicator and the day-of-week fixed effects. The motivation for this model is provided in Figure 1 and discussed in Section 2.1.

3. Results

Table 2 displays the first set of regression analysis results. Column (1) corresponds to the results from model (1), which indicate strong statistical evidence of a relationship between wind direction and $PM_{2.5}$ readings at the Avalon monitor (all of the ambient pollution data included in models (1) through (4) are gathered from the Avalon monitor which is in close proximity to the Shenango mill). That there is an association is intuitive: particularly strong sources of emissions will only affect air quality at the monitor for a particular range of wind directions. The fitted quadratic function bears an inverted U-shape with a maximum effect when the wind blows *from* 210° . Importantly, the partial effect contributes over $8 \text{ ug}/\text{m}^3$ at 210° . Note that the average $PM_{2.5}$ reading is $12.6 \text{ ug}/\text{m}^3$ at the Avalon monitor. The partial effect of wind direction comprises an important determinant of ambient $PM_{2.5}$.

Figure A.2. plots $PM_{2.5}$ levels against wind direction. The left-hand panel of the figure displays the raw hourly readings. The right side collapses the hourly data into averages by wind direction. From both plots it is clear that $PM_{2.5}$ levels are highest when the wind blows from the southwest. The raw hourly data show extremely high values (over $75 \text{ ug}/\text{m}^3$) are associated with the wind blowing from about 210° . This figure also indicates that the maximum $PM_{2.5}$ levels correspond to periods when the wind is blowing from the Shenango mill toward the Avalon monitor. That is, the vertical lines shown in both panels of Figure A.2 represent the bearing between the main smokestack at the Shenango mill and the monitor. The figure and the fitted quadratic between wind

direction and $PM_{2.5}$ suggest that the mill has an important role in dictating extreme pollution levels at the monitor.

Counterintuitively, the visual smoke variable is significantly *negatively* associated with ambient $PM_{2.5}$ readings; a one-pixel increase in smoke density is associated with a 0.001 $\mu\text{g}/\text{m}^3$ decrease in $PM_{2.5}$. The explanation for this result is a function of how the visual smoke data is gathered. Consider that the camera detects higher smoke readings when pixels in the field are most obscured. This occurs when the winds blow *across* the camera's field of vision; in this case, any actual particulate matter emitted by the plant is spread out across numerous pixels in the camera's field. Contrast this with a case in which winds blow *toward* the camera. Then any particulate emissions are less likely to be dispersed among multiple pixels thereby reducing the camera's smoke reading. While the precise location of the camera is not publicly known (the camera is located near a community members' home and exact location is withheld for privacy concerns), it is in the same general direction, relative to the mill, as the Avalon monitor. Thus, higher pollution periods tend to occur when the winds blow toward the camera.

Column (2) in Table 2 displays the results from model (2). Including the interaction term between smoke and wind direction does not appreciably alter the coefficients for wind direction. However, the smoke readings are no longer significantly associated with $PM_{2.5}$ levels – either directly or when interacted with wind direction.

Column (3) shows the fitted coefficients from model (3). This model includes one-hour lagged smoke readings along with contemporaneous smoke readings. Both measures

appear in the model directly and interacted with wind direction. As in models (1) and (2), wind direction is significantly related to PM_{2.5} levels; the coefficients are of the same sign and similar magnitude. In addition, none of the contemporaneous smoke controls are associated with PM_{2.5} readings. However, the lagged smoke measurements are significant determinants of ambient PM_{2.5}. A one-pixel increase in the direct measure of smoke is associated with a 0.005 ug/m³ decrease in PM_{2.5}. This is roughly five times larger than the effect of smoke estimated in model (1). The interaction between the one-hour lagged readings of smoke and wind direction suggests an inverted U-shape functional form with respect to the effect of these covariates on ambient PM_{2.5}. The partial effect maximizes at a wind direction of roughly 200°. The association between the smoke readings and the monitor readings for PM_{2.5} is positive, combining the partial effect of smoke through both the direct and wind interaction terms, for periods of time when the wind blows from 130° to 270°.

Model (4) includes a dummy variable corresponding to the intermittent power losses at the plant. This control is included because there were fires and copious smoke emissions at the facility during the outages. First, the aforementioned association between the smoke readings from the camera and ambient PM_{2.5} readings is essentially unchanged. Second, the results show a statistically significant relationship between the incident and PM_{2.5} levels. The direct incident variable is associated with a 5.2 ug/m³ reduction in PM_{2.5}. This seems counterintuitive since there were fires and copious smoke emissions at the facility during the outages. As such, it is also important to

examine the effects of the incident through wind direction. Combining the direct effect of the incident and the interaction terms (with wind direction) reveals that PM_{2.5} levels were higher during the incident when the wind was blowing from 50° (about northeast) through 300° (about northwest). The maximum effect occurred when the wind was blowing from the due south; the partial effect, conditional on this wind direction, was about 4 ug/m³. Thus, for the majority of realized wind directions, the power outages and associated fires increased ambient PM_{2.5}.

Table 3 tests for associations between ambient readings of SO₂ and H₂S and the visual smoke data. Our prior here is a finding of no association because both of these pollutants are gaseous and should be invisible to the camera's smoke detection algorithm. Columns (2) and (3) confirm this basic hypothesis. There is no evidence of a statistical association between the smoke readings produced by the camera and ambient readings of these gases. This serves as a useful placebo test.

3.1 Citizen Responses to the Camera

We begin by noting qualitative evidence suggesting a link between ACHD complaints and the activation of the camera and the concomitant livestreaming website. ACHD employees record summaries of caller comments, and, Figure 2 contains a selection of these comments that explicitly note the livestreaming channel in their calls. Such evidence motivates a more systematic assessment of this link.

Figure 3 presents the weekly count of total air quality complaints made to the ACHD in 2014 and 2015, with vertical lines indicating the first full week following the livestreaming start dates in 2014 and 2015 (the first and last vertical lines), as well as the week after the livestream temporarily stopped for over a month at the end of 2014 (the middle vertical line).

Table 4 reports regression results formally testing the association between all air pollution-related calls to ACHD and the presence of the camera and website. All results in Table 4 correspond to the negative binomial regressions, as previously mentioned. Column (1) corresponds to model (5), the most parsimonious specification including indicator variables for days during which the camera was active and days immediately following the camera's initial activation, or onset. We report a significant increase in call counts on days with the camera active. Specifically, there was an increase of one call every two days ($p < 0.01$) when the camera was active compared to days when the camera was not running. We find no evidence of an effect only on days when the camera was activated. (All further references to the indicator for the camera in Table 4 are for the indicator of days on which the camera was active.) Adding seasonal fixed effects, weather and pollution data, and controlling for time trends both reduces the magnitude of the effect and renders the camera active control insignificant. We do find evidence that on days following the announcement about the closure of the Shenango mill, there were about 1.4 fewer calls per day ($p < 0.01$). Column (3) adds interactions between the camera indicator and the pollution readings from monitoring stations. The

indicator for the camera being active now has a negative coefficient, but it remains statistically insignificant. The effect of the closure of the mill is similar to that reported in column (2). Finally, in column (4), we include day-of-week fixed effects and interactions of these terms with the camera-active control. The results from this model provide evidence consistent with Figure 1; the camera boosts the Monday peak in calls and reduces the tapering of calls in the rest of the workweek (relative to weekend call counts). Recall from Figure 1 that average call counts peaked on Mondays, and then declined throughout the remaining weekdays. The results in column (4) of Table 4 indicate that the camera is associated with an increase in this peak, by about one call every three days ($p < 0.01$). This is roughly a 20% increase in call counts over the Monday fixed effect. On the other weekdays, the camera corresponds to an increase in the weekday fixed effect by one call every two days ($p < 0.01$). This amounts to a 55% rise in call counts, relative to the weekday fixed effect. Additionally, in this specification the mill closure announcement indicator variable remains a significant determinant of call counts ($p < 0.01$), and the indicator variable for the days on which the local newspapers published articles about the camera increased call counts by about one call every two days ($p < 0.10$). The indicator for the day on which the community meeting was held is associated with a reduction of about one call every two days ($p < 0.10$). Using the specification in column (4), the linear combination of either of the day-of-week interaction terms with the camera active variable is not significant at conventional levels. (Tables A2 through A4 in the appendix display the full econometric results corresponding to tables 4, 5, and 6.)

Table 5 focuses on calls made to the ACHD that mention, or clearly refer to, the Shenango mill. Figure 4 plots the weekly Shenango call counts against time. The vertical lines correspond to when the camera was activated, stopped, and reactivated (as in Figure 3). Akin to the results in Table 4, in the parsimonious model shown in column (1) we report a significant association between the indicator variable corresponding to days on which the camera was active ($\alpha = 0.01$). The estimated coefficient suggests calls increase by about one call every three days. However, adding seasonal fixed effects, time trends, and controls for pollution and weather eliminate this effect. The mill closure announcement variable suggests that there were roughly two fewer calls per day about Shenango after the announcement.

In columns (2), (3), and (4) the estimated coefficient on the camera active control is negative, though, not generally significant. In column (4), there is weak evidence that the camera-onset variable is associated with an increase in calls about Shenango of about one per day. Finally, we find no evidence of an effect of the camera when interacted with the day-of-week fixed effects. The indicator variable for the day on which the community meeting was held is associated with a reduction of about one call per day ($p < 0.05$). Using the specification in column (4), the linear combination of either of the day-of-week interaction terms with the camera active variable is marginally significant ($p < 0.10$) and negative; the combined effect of the camera is a reduction of between four and five calls per day.

At first, this result seems counterintuitive. Why would the camera be associated with lower call counts? One candidate explanation is that the well-publicized livestream affected emissions produced by the mill. While we cannot test this directly without emission readings both before and after the onset of the camera, we can leverage the hourly PM_{2.5} observations and wind direction data from the Avalon monitor. Shenango is situated under one-half mile and 210 degrees from the monitor. One way to test whether emissions released by the plant changed after the camera was installed is to assess ambient PM_{2.5} levels at the Avalon monitor. To do so, we restrict the sample to those hours during which the wind blew from four different direction ranges, all centered at 210 degrees. We then conduct a t-test comparing the hourly PM_{2.5} readings before and after the camera was activated. These results are shown in table A1 in the appendix. Beginning with the widest direction band between 165 and 255 degrees, the test rejects the null hypothesis of equal pre-and-post-camera means at ($p < 0.001$). The absolute difference is 14.2/m³ before the camera was launched and 13.2 ug/m³ after the camera was activated. This amounts to a 6.7% reduction in hourly PM_{2.5} readings. We find similar results for the specifications using 175 to 245 degrees and 195 to 225 degrees, though the significance of the rejection of the null hypothesis weakens. Employing the 200 to 220 specification, we find a nearly 9% reduction in ambient PM_{2.5}; here we reject the null at ($p < 0.05$). Finally, when we restrict observations to those in which the wind was blowing from 210 degrees plus or minus just five degrees, the mean difference is much larger: nearly 20 percent ($p < 0.01$).

One interpretation of the negative effect that the camera has on Shenango-related call counts is that the camera caused a change in behavior at the mill, which reduced emissions, and in turn, ambient concentrations in the community from which most of these calls emanate. That is, if calls are associated with visual emissions and ambient concentrations of $PM_{2.5}$, then the significant reductions after the camera was installed may explain this finding. The results reported in table A1, especially those for the 205 – 215 degree specification appear to support this argument.

Table 6 displays the results of the regression analyses that employ calls pertaining to steel mills. Figure 5 plots the weekly call counts against time. The vertical lines correspond to when the camera was activated, stopped, and reactivated (as in Figures 3 and 4). A dramatic uptick in call counts following a few weeks after the final activation of the camera is evident in the figure. This is reflected in column (1) of Table 6, which reveals a large and significant relationship between the camera being active and call counts ($p < 0.01$). The effect is over three calls every two days. Much like the results reported in Tables 4 and 5, however, this effect is not robust to the inclusion of seasonal fixed effects, the time trend, and controls for weather and pollution conditions. In columns (2) and (3), neither controls for the announcement of the closure of the Shenango mill nor the days on which articles about the camera were published are significantly associated with call counts pertaining to steel mills. In column (4), the control for days on which articles about the camera appeared in local newspapers is significantly positively related to call counts about steel mills, while the announcement

of the closure of the Shenango mill had no effect on call counts. Notably, we report significant evidence of an effect of the camera on call counts when interacted with the day-of-week fixed effects. In particular, there was an estimated one more call per day on Mondays relative to weekends when the camera was active than when the camera was inactive ($p < 0.10$). An effect of three calls every two days is detected for other weekdays ($p < 0.01$). To put this into perspective, this later effect boosts the later weekday fixed effect by 10% relative to the pre-camera period, resulting in a noticeable reduction in the tapering of calls throughout the work week for the camera on period (see the bottom panel of Figure 1 in which steel-related calls rise at the end of the week in the post camera period). Taken in total, these findings suggest that the camera, working through the interactions with day-of-week fixed effects, had a persistent (if heterogeneous) effect on the propensity of citizens to call ACHD about air pollution produced by steel mills.

3.2 Spatial Analysis of Citizen Complaints About Shenango

Investigation of the location of origin for ACHD complaints reveals further suggestive evidence of a possible role of the information provided by the Shenango livestream in ACHD complaint generation. Whenever possible, ACHD records the zip code of all complainants, allowing us to exploit geographical variation in the public response to Shenango pollution. Importantly, 79% of all Shenango-related complaints come from one zip code: that which contains the Avalon neighborhood and the camera. This zip code lies directly to the north east of the Shenango coke plant. This area is downwind from Shenango when the wind blows from the most frequent wind direction. The

fraction of all Shenango complaints coming from this zip code is the same both before and during the Shenango channel livestream broadcast. In addition to being downwind from the mill conditional on the most common wind direction, the topography of this neighborhood is such that many residents have a direct view (unassisted by the camera livestream) of the Shenango plant itself and its visible pollution. Therefore, it is difficult to empirically disentangle an effect of direct observation versus information provided by the camera on calls from this area.

On the other side of the river (the southwest bank relative to the Shenango plant) the topography leaves the Shenango coke plant largely out of view to most residential neighborhoods. Residents in the two zip codes that lie to the southwest of the mill do not generally have a direct line of sight to the Shenango-coke plant. This suggests that the online camera may comprise a greater shock to the set of information commonly accessible to residents of these zip codes relative to those in neighborhoods from which the mill can be seen. Visual evidence of pollution provided by the camera may be more important in assisting residents from the southwest zip codes in the attribution of ambient pollutants to a source. Notably, calls from these two southwest bank zip codes increase dramatically from two calls in the period in 2014 and 2015 before the livestream to 16 calls during the same period when the livestream was operational. (The periods with and without the camera are almost exactly the same duration). A ranksum test confirms that there is indeed a significant difference in the share of Shenango-

related calls to ACHD from these southwest bank zip codes between the periods in which the livestream is active and when it is not (p-value of 0.0047).

Of course, other factors simultaneous to the period of livestream activity could be causing an increase in calls from the southwest bank. The obvious concern is that wind patterns may happen to direct the Shenango smoke to the southwest bank more often in the period during which the camera was active. We test for this non-parametrically by comparing the distribution of wind direction when the camera was active and when it was inactive. Figure 6 and Figure 7 present, respectively, the cumulative distribution functions (cdfs) and kernel densities of the hourly-normalized wind direction distribution⁶. The cdfs and densities in Figure 6 and Figure 7 reflect two periods: when the camera was inactive and when it was active during 2014 and 2015 up until the announcement of the Shenango plant closure. As can be seen, there is little difference in the camera-active and camera-inactive periods. A Kolmogorov-Smirnov test with the null hypothesis that the distribution of wind points more in the direction of the southwest bank when the camera is on than when it is off rejects the null ($p < 0.000$). This suggests no support for this alternative explanation for the increase in southwest bank calls during the livestreaming period.

⁶ The normalization is such that 0 represents the direction from which the wind would come in order to make the average of the geographic centers of the south bank zip codes directly down stream from the Shenango coke mill, with 1 degree representing a wind direction coming from a degree away (on either side) from this point.

5. Conclusions

Most environmental policy assumes the form of standards and enforcement. However, because both monitoring and enforcement are expensive, these efforts encompass just a sample of polluters. The fact that both enforcement and monitoring are not comprehensive motivates the use of disclosure laws. This study explores a new form of pollution disclosure: real-time visual evidence of emissions provided on a free, public website. Real-time broadcasts of emissions differ in important ways from extant rules such as TRI and the GHGRP. First, gathering visual evidence does not rely on firm reporting. In fact, the approach used in this study completely obviates the firm's internal pollution tracking efforts, and, even in the face of a disinterested or overburdened governmental regulatory authority, the approach studied opens avenues for polluter accountability before the public. Second, the imagery provides graphic evidence of transgressions by firms. This is likely to engender a very different response among citizens relative to emissions data disclosed in tabular form.

We develop a new dataset comprised of daily call counts to the local air quality regulator that we employ to test whether this new form of disclosure affects the nature and frequency of calls that citizens make to the regulator. Ultimately, the paper seeks to evaluate whether visual evidence offers a viable complement to traditional approaches to managing pollution.

The literature focusing on how individuals respond to risk messaging guides our empirical strategy. That is, one factor that dictates whether people respond

systematically or emotionally to information about risk is the amount of prior knowledge they have about the subject (Averbeck, Jones, and Robertson, 2011). Because the preponderance of calls focusing on the Shenango mill originate from callers in the same zip code, we claim that their prior knowledge about emissions is high. In contrast, citizens in other zip codes are less likely to have knowledge about the visually evident emissions from Shenango. After testing for an association between the camera being active and all air pollution-related calls, we test the prior knowledge hypothesis by subdividing the sample of calls into those specifically about the Shenango Mill and those that are about the other large industrial source of air pollution. Specifically, we employ a subsample of calls just about steel manufacturing facilities. There are no such facilities in the same zip code as Shenango and calls about this source type are the second largest industrial category of complaints. Because of less prior knowledge held by these callers, we expect to see a greater response to the camera imagery among this subset of calls than among calls about Shenango.

We detect evidence of an association between the daily count of all calls pertaining to air pollution in the ACHD's database and the presence of the camera. In our preferred specification, we find evidence that interactions between the active camera indicator and day of the week fixed effects are significant determinants of call counts. The interaction with the Monday fixed effect suggests an increase, above and beyond the pre-camera period, of 0.3 calls per day ($p < 0.01$) on Mondays (relative to the count of calls on weekend days). The interaction between the camera control and other

weekdays indicates that the camera produces an increase twice as large ($p < 0.01$). The models that feature complaints about Shenango, the monitored facility, suggest that the camera had no clear effect on daily call counts. Our final set of tests detects evidence of an effect of the camera on the count of calls about steel mills through the interaction with day-of-week controls. For example, the interaction term with the Monday control suggests that the camera was associated with an increase of 1 call per day, relative to weekend days during the period when the camera was active ($p < 0.01$).

Our results broadly comport with the prior knowledge hypothesis. The frequency of calls specifically about Shenango is not affected by the onset of the camera. People who call the ACHD about the mill tend to live near the mill. They routinely observe, and are exposed to, its discharges. The camera provides limited information. In contrast, when we limit our analysis to calls about another class of major point source of air pollution we find an enduring effect of the camera. Since these calls originate in areas more distant from Shenango we posit that callers have less knowledge about emissions from the mill. The prior knowledge hypothesis predicts that these communities react more emotionally to risk messaging. Indeed, we find evidence of such responses. The uptick in calls for steel-related complaints also embodies aspects of the affect heuristic; particularly the formulation put forth by Shiller (2017), in which people responding emotionally to stimuli often apply it to unrelated, or not directly related, circumstances. A citizen living near a steel mill reads about the Shenango camera, visits the website, sees graphic evidence of air pollution, reacts emotionally, and then when they observe

emissions from the facility nearby where they live, they are more prone to complain.

This is a reaction based on the affect heuristic and it may help to explain the enhanced call counts to the ACHD about steel manufacturing plants.

There are a number of ways to communicate the risks associated with air pollution exposure. Regulatory agencies provide current measurements of pollution and characterize associated risks. The USEPA publishes an air quality index that employs categories of risk (USEPA, 2017). Alternatively, activists and other stakeholders communicate risk through personal experience or narrative-based storytelling in popular media outlets (Pittsburgh Post-Gazette, 2015a; 2015b). Livestreaming images of emissions is an alternative tack to conveying risk to the public. Publishing the live imagery of emissions is more tangible, more emotionally charged, than either approach described above.

Whether due to the prior knowledge hypothesis or the affect heuristic (or a combination of the two) the fact that the Shenango camera appears to induce an increase in calls to the ACHD about pollution from other sources suggests that visual evidence may provide a valuable tool in boosting citizens' willingness to engage with regulators about pollution. In an era in which traditional monitoring and enforcement efforts may be on the decline, this new tool may be an especially important complement to traditional management of pollution.

This paper suggests further research on the efficacy of disclosure through visual evidence in potentially numerous contexts. For example, researchers could test whether

visual disclosure of other forms of pollution affect citizen engagement with regulators. These might include water or solid waste pollution. Future studies could design more tightly controlled experiments in which the type of visual evidence differs across random samples of individuals to glean what aspects of visual evidence is most effective in triggering a response. Further, future work could explore whether citizens' responses are sensitive to environmental justice issues: does visual evidence about pollution in distressed communities engender a different response than that provided in more affluent locales?

Tables

Table 1: Summary Statistics

Variable	mean (std. dev.)	min	Max
PM _{2.5} (ug/m ³)	15.28 (9.06)	2.6	63.8
SO ₂ (ppb)	21.77 (22.99)	0.2	244
O ₃ (ppb)	41.42 (14.78)	5	84.0
H ₂ S (ppm)	0.000 (0.001)	0	0.011
Temp. (°C)	17.27 (6.75)	-4.6	30.6
Pressure (mm Hg)	742.29 (4.19)	727.7	754.6
Smoke (pixels)	71.48 (288.18)	0	2,666
Wind Speed (mph)	3.82 (1.91)	0	13.1
Wind Direction (degrees)	204.36 (84.68)	0	359
All Calls per Day	3.52 (4.07)	0	31
Shenango Calls per Day	1.33 (2.32)	0	16
Steel Mill Calls per Day	0.10 (0.41)	0	5

Table 2. The Determinants of Pollution Levels at the Avalon Monitor

Covariates	(1)	(2)	(3)	(4)
Wind Direction	0.0818*** (0.00702)	0.0817*** (0.00720)	0.0796*** (0.00727)	0.0710*** (0.00728)
(Wind Direction) ²	-0.000195*** (1.76e-05)	-0.000196*** (1.80e-05)	-0.000191*** (1.82e-05)	-0.000167*** (1.82e-05)
Smoke	-0.00102** (0.000431)	-0.00128 (0.00125)	-0.00112 (0.00116)	-0.00137 (0.00117)
Smoke _{t-1}			-0.00540** (0.00228)	-0.00567** (0.00227)
Smoke x Wind Direction		-6.73e-06 (1.86e-05)	-8.78e-06 (1.79e-05)	-6.03e-06 (1.80e-05)
(Smoke x Wind Direction) ²		3.68e-08 (5.80e-08)	4.26e-08 (5.64e-08)	3.70e-08 (5.68e-08)
Smoke x Wind Direction _{t-1}			6.14e-05** (2.85e-05)	6.45e-05** (2.83e-05)
(Smoke x Wind Direction) ² _{t-1}			-1.54e-07** (7.46e-08)	-1.60e-07** (7.38e-08)
Incident				-5.202* (2.857)
Incident x Wind Direction				0.0967*** (0.0300)
(Incident x Wind Direction) ²				-0.000264*** (7.24e-05)
Constant	4,744* (2,533)	4,790* (2,537)	4,796* (2,549)	5,167** (2,543)
N	2,836	2,836	2,836	2,836
R ²	0.400	0.401	0.402	0.410

Dependent variable: Hourly PM_{2.5}
 Robust standard errors in parentheses
 *** p<0.01, ** p<0.05, * p<0.1

Table 3: The Determinants of PM_{2.5}, SO₂, and H₂S Levels at the Avalon Monitor

Covariates	(1) PM _{2.5}	(2) SO ₂	(3) H ₂ S
Wind Direction	0.0706*** (0.00726)	8.11e-06*** (2.57e-06)	2.44e-06*** (8.08e-07)
(Wind Direction) ²	-0.000167*** (1.82e-05)	-2.23e-08*** (6.37e-09)	-5.88e-09*** (1.96e-09)
Smoke	-0.00147 (0.00120)	1.98e-07 (6.62e-07)	9.44e-09 (1.51e-07)
Smoke _{t-1}	-0.00575** (0.00228)	-1.17e-06 (1.39e-06)	7.87e-08 (1.99e-07)
(Smoke x Wind Direction)	-3.63e-06 (1.82e-05)	1.90e-09 (1.17e-08)	-7.47e-10 (2.59e-09)
(Smoke x Wind Direction) ²	2.91e-08 (5.71e-08)	-0 (0)	0 (0)
(Smoke x Wind Direction) _{t-1}	6.62e-05** (2.84e-05)	1.46e-08 (1.97e-08)	-1.28e-09 (2.38e-09)
(Smoke x Wind Direction) _{t-1} ²	-1.65e-07** (7.41e-08)	-0 (0)	0 (0)
Incident	-5.201* (2.843)	0.00160*** (0.000601)	1.28e-05 (0.000170)
(Incident x Wind Direction)	0.0937*** (0.0299)	-2.62e-05*** (6.78e-06)	1.00e-06 (2.14e-06)
(Incident x Wind Direction) ²	-0.000253*** (7.22e-05)	6.18e-08*** (1.68e-08)	-4.64e-10 (5.22e-09)
Constant	6,311** (2,491)	1.570 (1.687)	-0.786* (0.406)
N	2,836	2,836	2,836
Adj. R ²	0.409	0.362	0.386

Dependent variables shown in top row of table.

Robust standard errors in parentheses

*** p<0.01, ** p<0.05, * p<0.1

Table 4: Determinants of All Complaints to ACHD

Covariates	(1)	(2)	(3)	(4)
Camera	-0.0250	-0.0759	-0.0302	-0.0488
Onset	(0.201)	(0.224)	(0.226)	(0.175)
Camera	0.462***	0.233	-0.944	-0.827
Active	(0.0823)	(0.247)	(1.409)	(1.171)
Closure		-1.390***	-1.366***	-1.254***
		(0.285)	(0.286)	(0.246)
Article		0.654	0.765	0.662***
		(0.519)	(0.540)	(0.148)
Camera x				0.310**
Monday				(0.143)
Camera x				0.582***
Weekday				(0.0893)
Monday				1.648***
				(0.100)
Weekday				1.060***
				(0.0639)
Constant	1.009***	-0.419	0.00154	-1.421
	(0.0565)	(0.708)	(1.084)	(0.892)
Season Fixed		X	X	X
Effects				
Cubic Time		X	X	X
Trend				
Weather		X	X	X
Controls				
Monitor		X	X	X
Pollution				
Camera x			X	X
Monitor Pollution				
Ln(alpha)	-0.526***	-0.698***	-0.710***	-1.429***
	(0.0652)	(0.0640)	(0.0634)	(0.104)
N	730	727	727	727

Standard errors in parentheses: * p<0.10, ** p<0.05, *** p<0.01

Dependent Variable is daily count of all calls to ACHD pertaining to air pollution.

Table 5: Determinants of Shenango Complaints to ACHD

Covariates	(1)	(2)	(3)	(4)
Camera	0.337	0.519	0.653	0.704*
Onset	(0.313)	(0.432)	(0.445)	(0.392)
Camera	0.280**	-0.184	-4.700*	-3.718
Active	(0.131)	(0.331)	(2.456)	(2.491)
Closure		-1.992***	-1.821***	-1.809***
		(0.562)	(0.577)	(0.553)
Article		0.498	0.623	0.250
		(0.608)	(0.612)	(0.298)
Camera x				-0.706
Monday				(0.893)
Camera x				-0.249
Weekday				(0.886)
Monday				5.099***
				(0.703)
Weekday				4.236***
				(0.697)
Constant	0.130	-4.605***	-2.631	-7.899***
	(0.0990)	(1.288)	(1.869)	(1.854)
Season Fixed		X	X	X
Effects				
Cubic Time		X	X	X
Trend				
Weather		X	X	X
Controls				
Monitor		X	X	X
Pollution				
Camera x			X	X
Monitor Pollution				
Ln(alpha)	0.882***	0.666***	0.657***	-0.188
	(0.0905)	(0.0970)	(0.0967)	(0.131)
N	730	727	727	727

Standard errors in parentheses. * $p < 0.10$, ** $p < 0.05$, *** $p < 0.01$
 Dependent Variable is daily count of Shenango calls to ACHD pertaining to air pollution.

Table 6: Determinants of Steel Complaints to ACHD

Covariates	(1)	(2)	(3)	(4)
Camera	-1.096	0.744	1.088	1.116
Onset	(0.770)	(0.770)	(0.785)	(0.807)
Camera	1.738***	-0.839	-1.513	-1.749
Active	(0.302)	(0.957)	(6.400)	(5.814)
Closure		0.328	0.445	0.432
		(0.761)	(0.786)	(0.726)
Article		0.675	0.831	0.923*
		(0.838)	(0.890)	(0.493)
Camera x				1.026*
Monday				(0.536)
Camera x				1.530***
Weekday				(0.487)
Monday				16.18***
				(0.461)
Weekday				15.45***
				(0.343)
Constant	-3.064***	-7.361**	-7.408	-23.72***
	(0.265)	(3.285)	(5.815)	(5.327)
Season Fixed		X	X	X
Effects				
Cubic Time		X	X	X
Trend				
Weather		X	X	X
Controls				
Monitor		X	X	X
Pollution				
Camera x			X	X
Monitor Pollution				
Ln(alpha)	1.025***	0.625*	0.517	-0.0739
	(0.301)	(0.349)	(0.350)	(0.420)
N	730	727	727	727

Standard errors in parentheses. * p<0.10, ** p<0.05, *** p<0.01

Dependent Variable is daily count of Steel Manufacturing calls to ACHD pertaining to air pollution.

Figures

Figure 1. Average ACHD Complaints by Day of Week with Camera On and Off

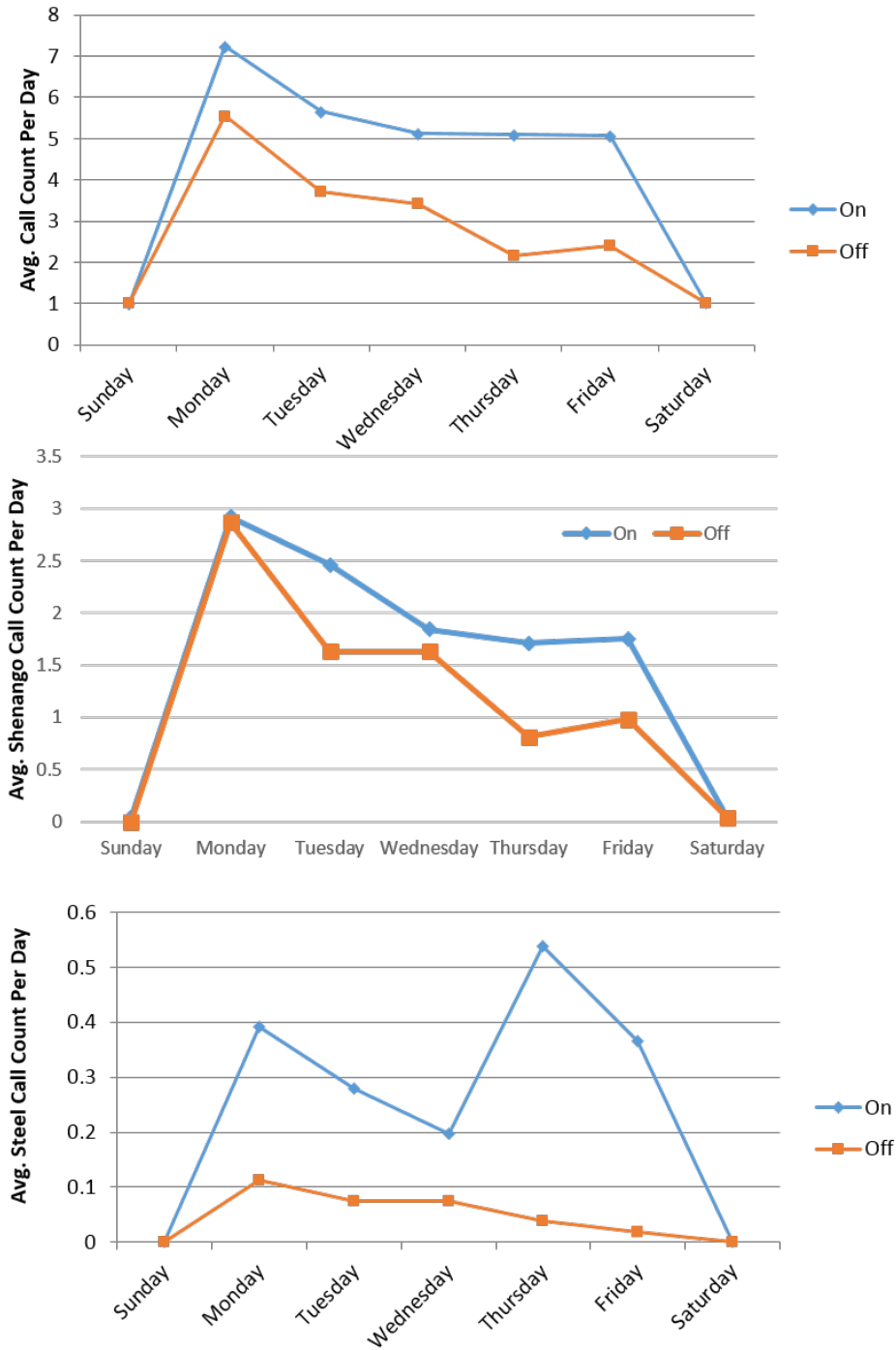


Figure 2: Text from ACHD Call Notes

Sample Complaint 1: ACCORDING TO **SHENANGOCHANNEL.ORG**, THE PM 2.5 LEVEL AT ABOUT 9 AM ON 11/2/15 WAS 47UG/M³ IN AVALON! THE AIR SMELLS HORRIBLE TODAY.

Sample Complaint 2: SHENANGO COKE WORKS - STRONG SMELL OF TAR AND GAS. CALLER SAID TO **CHECK THE CAMERA FEED** FOR 8:45 PM. THERE WERE 50 FT+ GAS FLARES SHOOTING OUT.

Sample Complaint 3: THE HOUSE REEKED OF CHEMICALS WHEN HE WOKE UP TODAY. WENT OUT FOR A WALK AT 6:20 AND THE BURNT, INDUSTRIAL SMELL WAS IN THE AIR. IT WAS STILL STINKY AT 7 AM WHEN HE GOT BACK. ACCORDING TO **HTTP://SHENANGOCHANNEL.ORG** THE WIND HAS BEEN BLOWING FROM THE WEST (AKA SHENANGO) ALL NIGHT.

Sample Complaint 4: WHAT THE HELL IS GOING ON AT SHENANGO? DESPITE YOUR RIDICULOUS AND USELESS CONSENT AGREEMENTS, THINGS ARE GETTING WORSE! WHAT THE HELL HAPPENED LAST NIGHT AT AROUND 7? **IT'S ON VIDEO:** BILLOWING BLACK SMOKE FOR SOME TIME, THE EMERGENCY FLARE SHOOTING UP AT LEAST 20 FEET! I WANT ANSWERS! THIS IS RIDICULOUS! I THINK THE NEWS NEEDS TO BE INFORMED AND THEN THE WHOLE DAMN CITY CAN SEE THE NEGLECT OF ACHD AND HOW YOU LET INDUSTRY POISON AND DESTROY THE LIVES OF THOUSANDS OF RESIDENTS. I HOPE YOUR FAMILIES ARE ALL CHIKING LIKE MINE DOES! HISTORY IS A HARSH JUDGE, THOSE WHO ARE COMPLICIT ARE AS GUILTY AS THE PERPETRATORS AND EVERY MEMBER OF ACHD WHO DOES NOT DO SOMETHING ABOUT THIS IS GUILTY!

Figure 3: Weekly Count of All Calls to the ACHD in 2014 and 2015

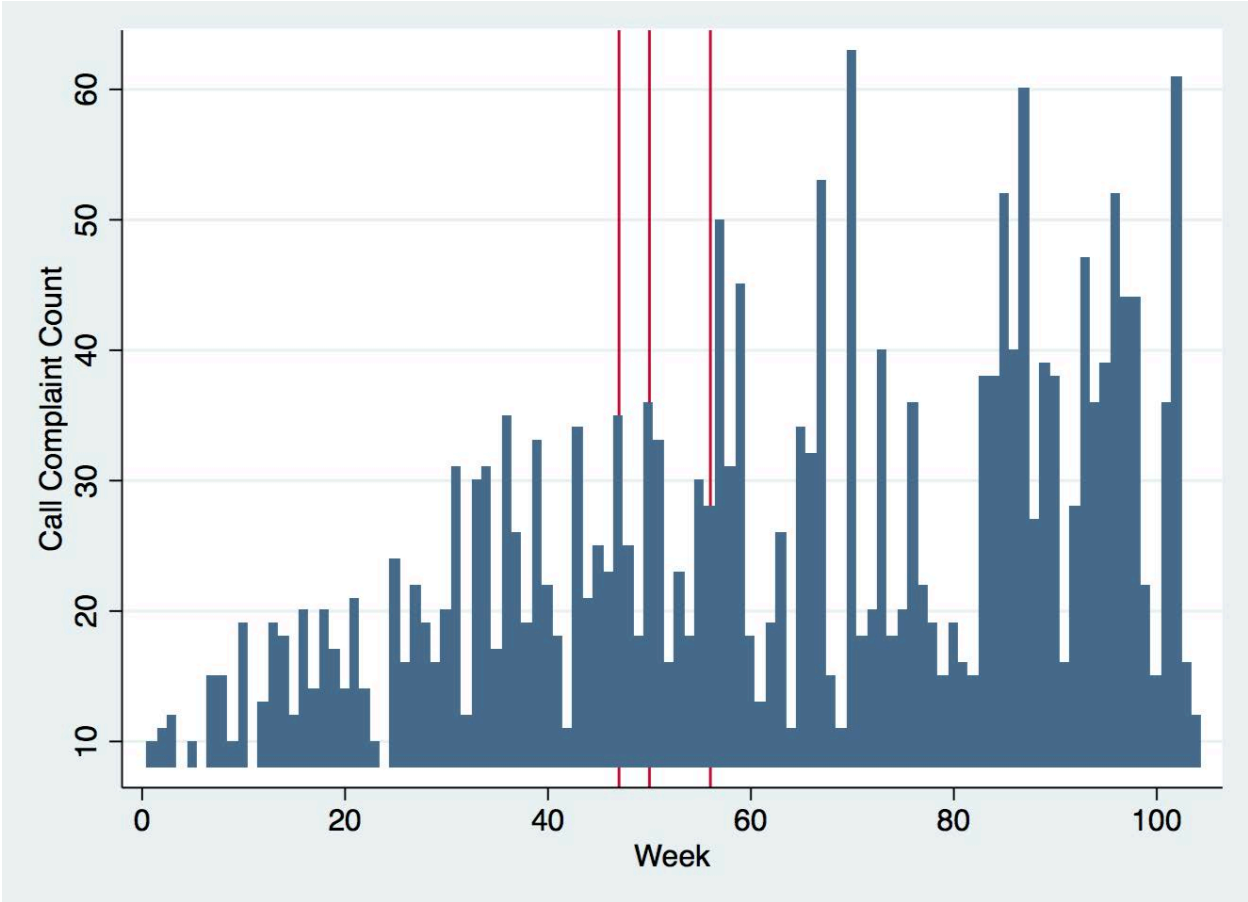


Figure 4: Weekly Count of Shenango-Related Calls to the ACHD in 2014 and 2015

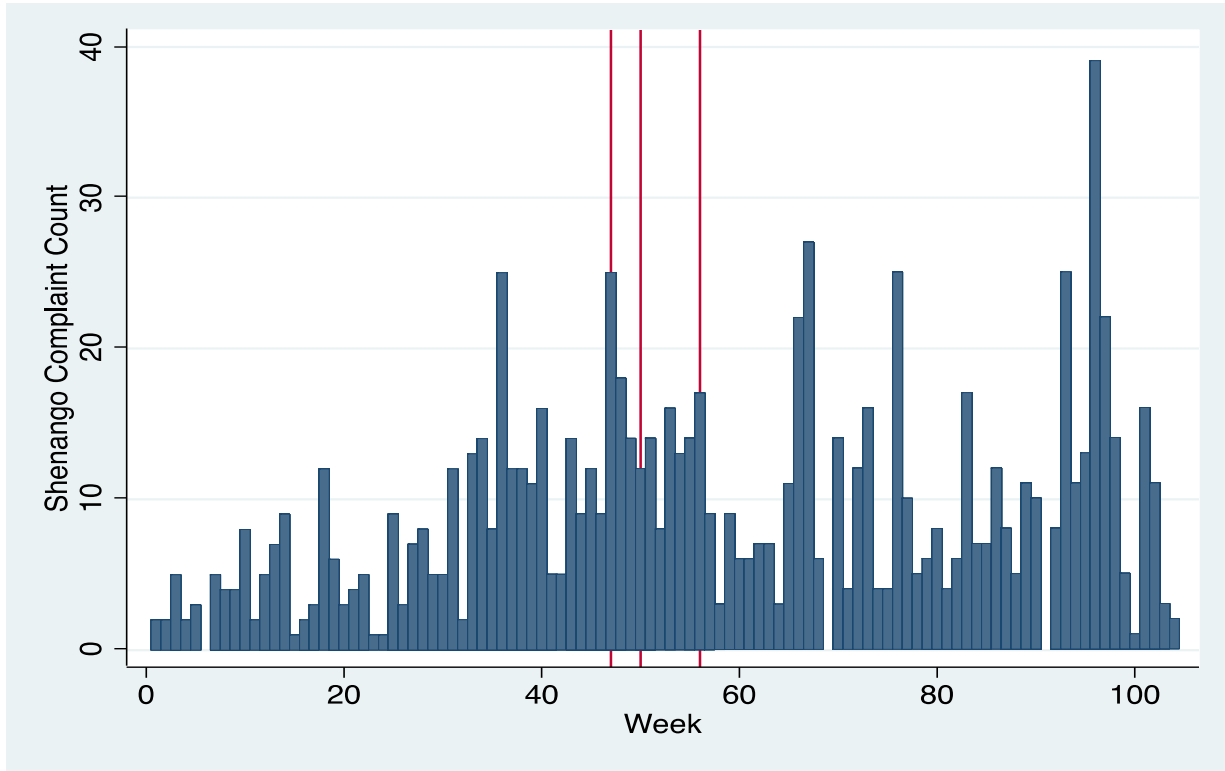


Figure 5: Weekly Count of Steel Pollution-Related Calls to the ACHD in 2014 and 2015

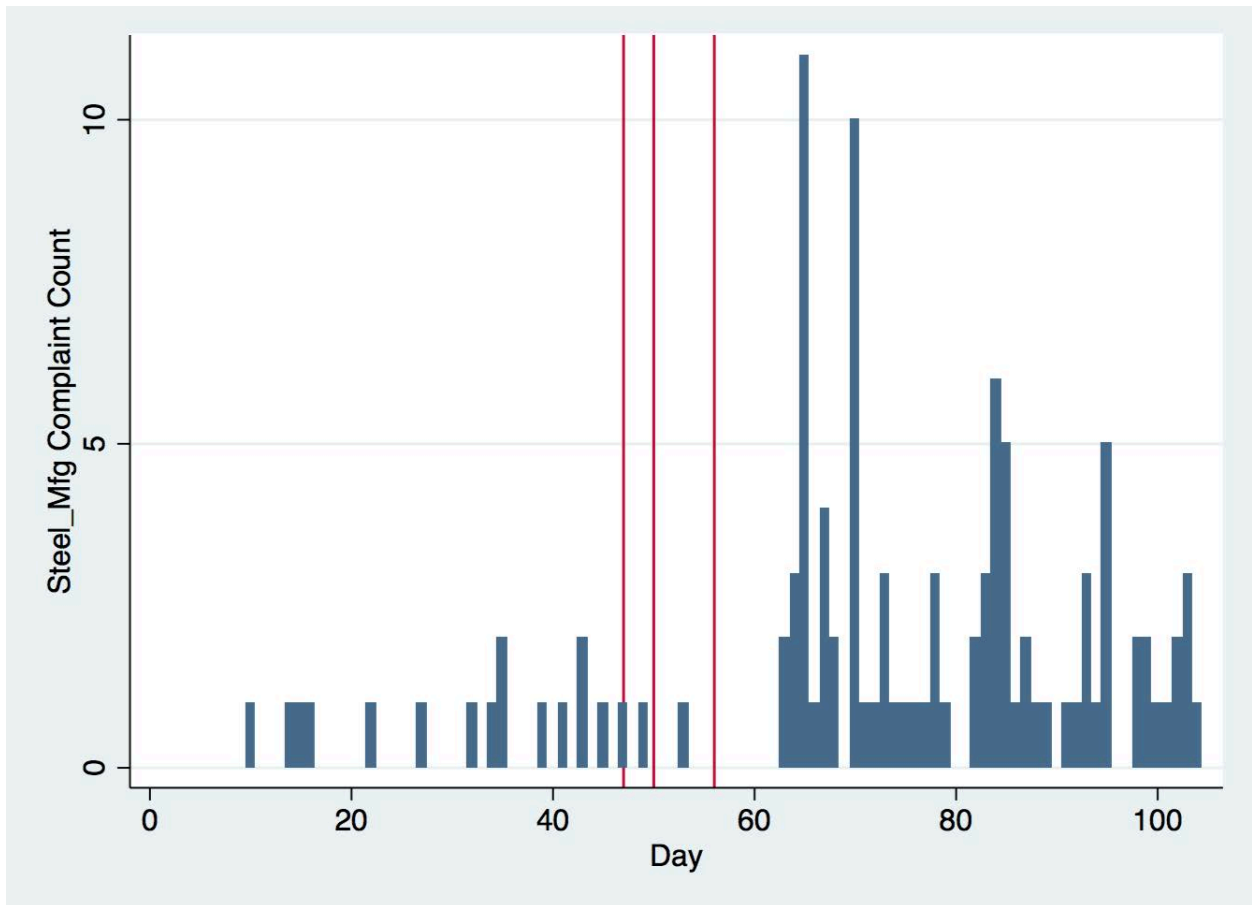


Figure 6: CDF of Normalized Wind Distribution Hourly Readings (Livestream Camera on and Off)

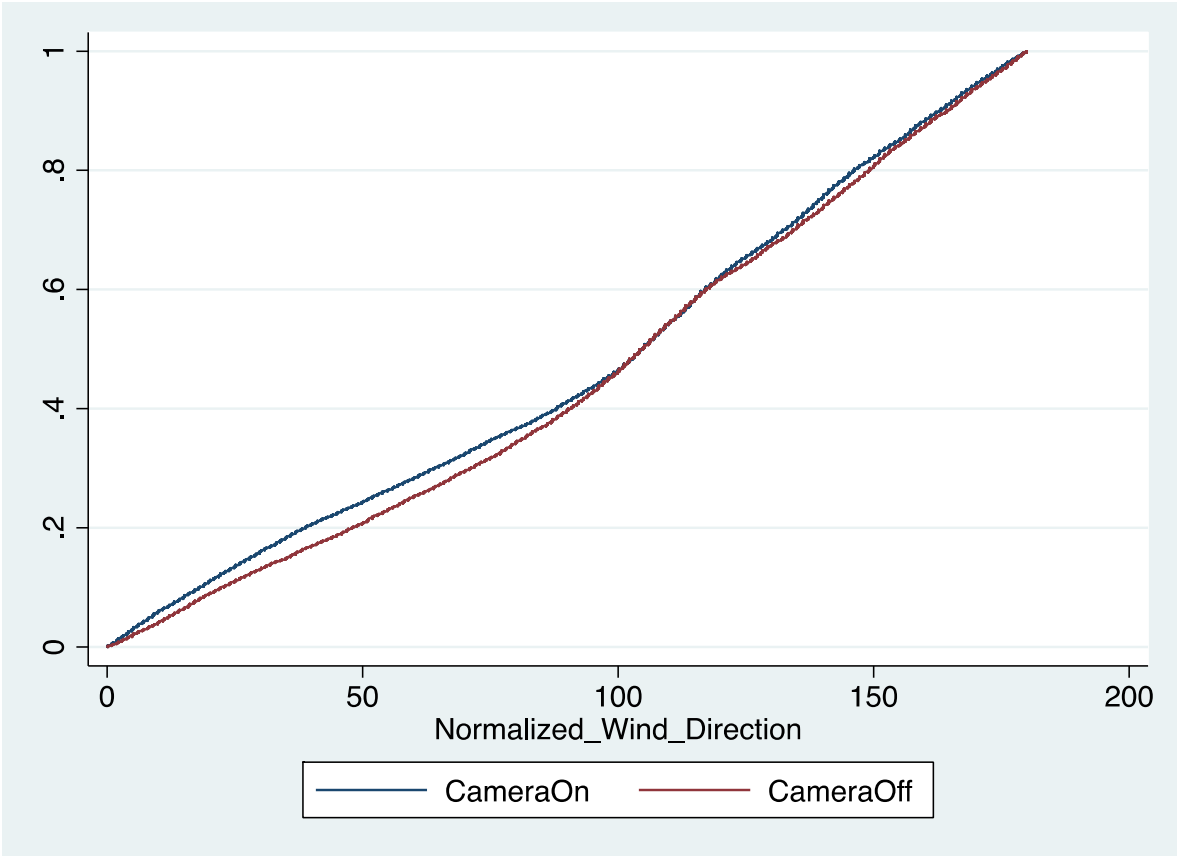
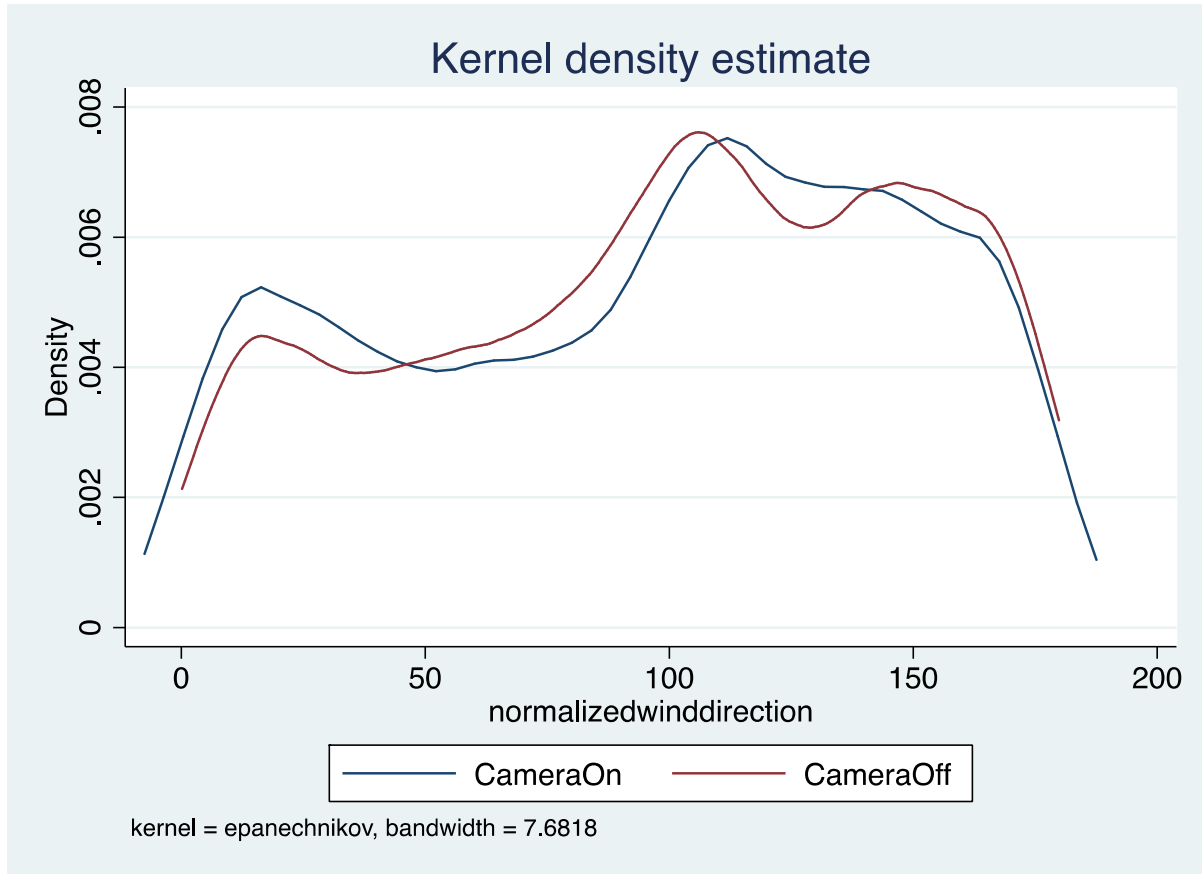


Figure 7: Kernel Density of Normalized Wind Distribution Hourly Readings

(Livestream Camera on and Off)



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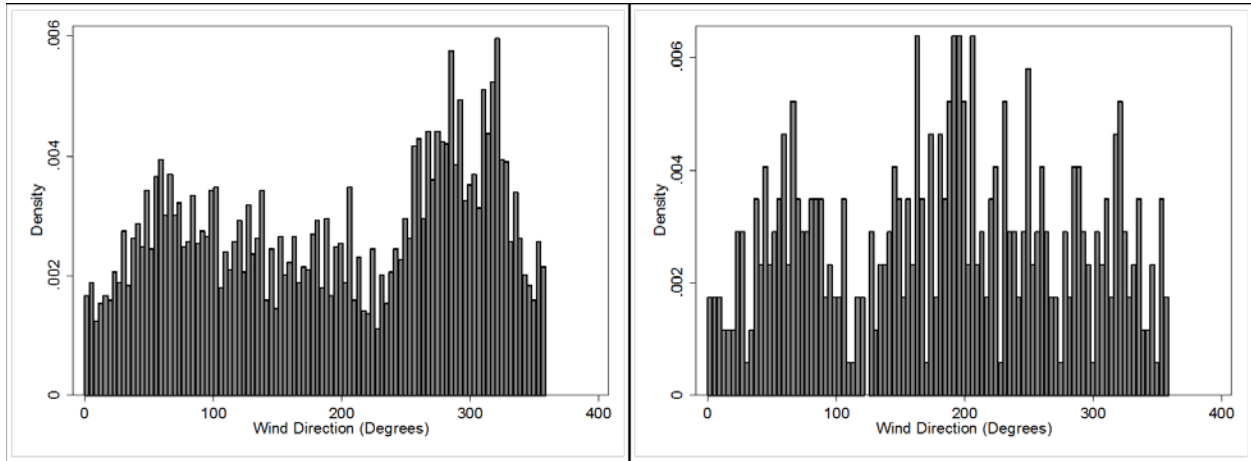
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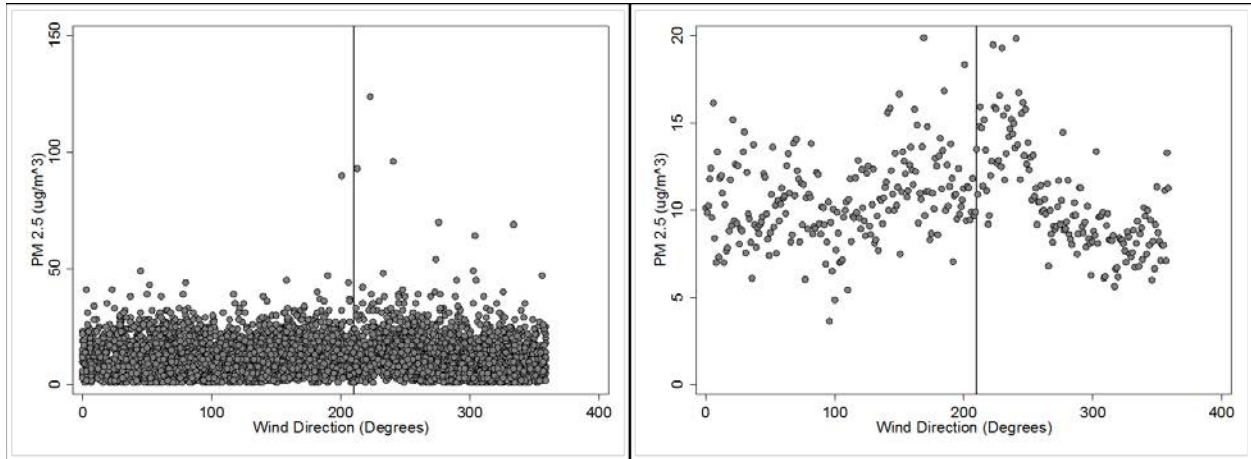
Appendix.

Figure A.1. Histograms of Wind Direction at Avalon Monitor.



Left Panel: Hourly wind direction from March, 2015 through November, 2015. Right Panel: Hourly wind direction during spring 2015 power outage.

Figure A.2. PM_{2.5} and Wind Direction.



Left Panel: Hourly PM_{2.5} Observations. Right Panel: Average PM_{2.5} Readings.

Vertical line = orientation of main stack at Shenango and Avalon monitor.

Figure A.3: Smoke Detection Algorithm.

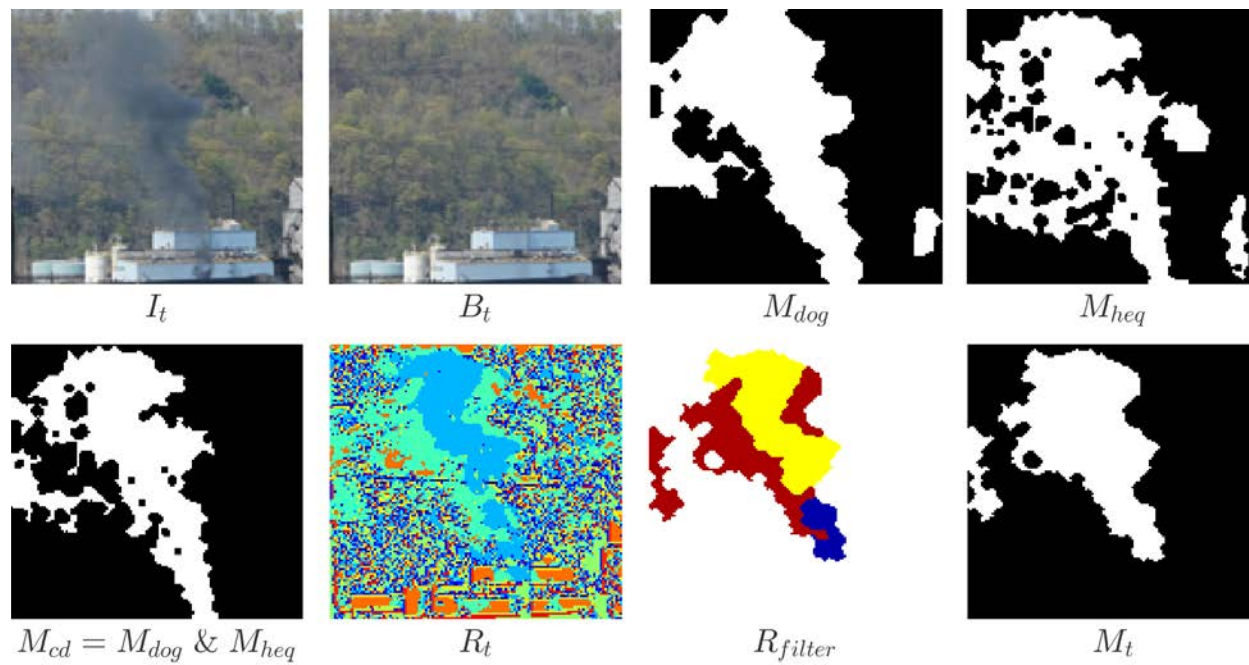


Figure A.4: Sample of Time Lapse View and Data Depiction.

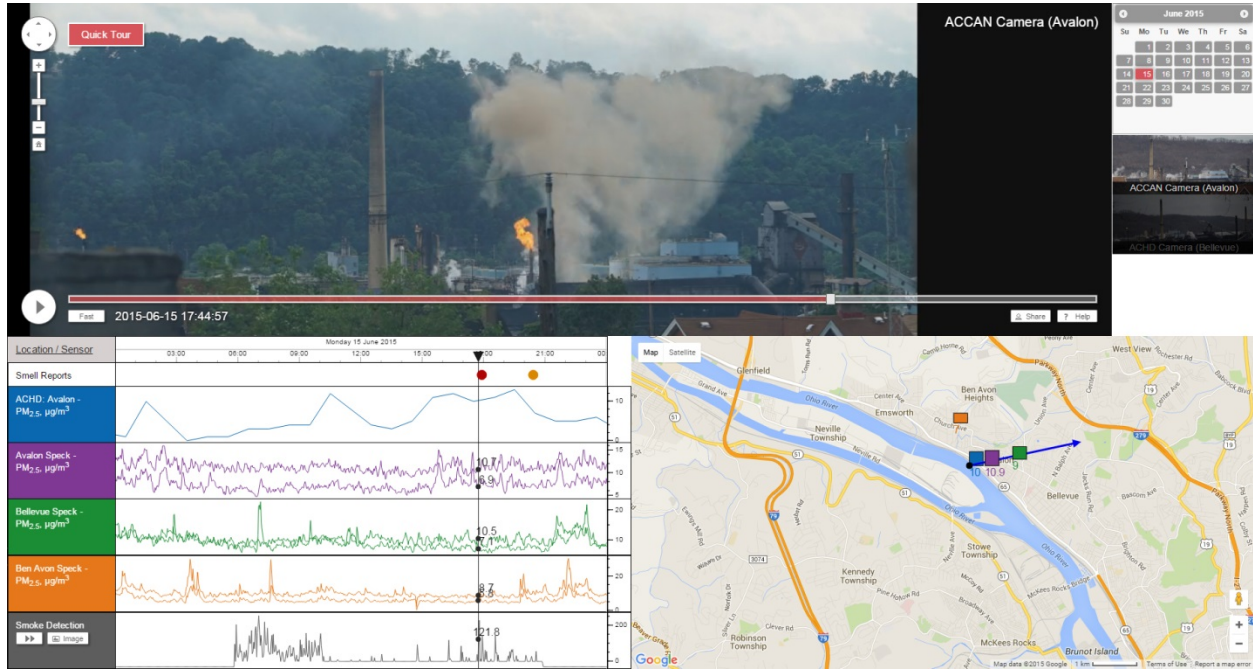


Figure A.5: Smoke Detection Results for May 2, 2015.

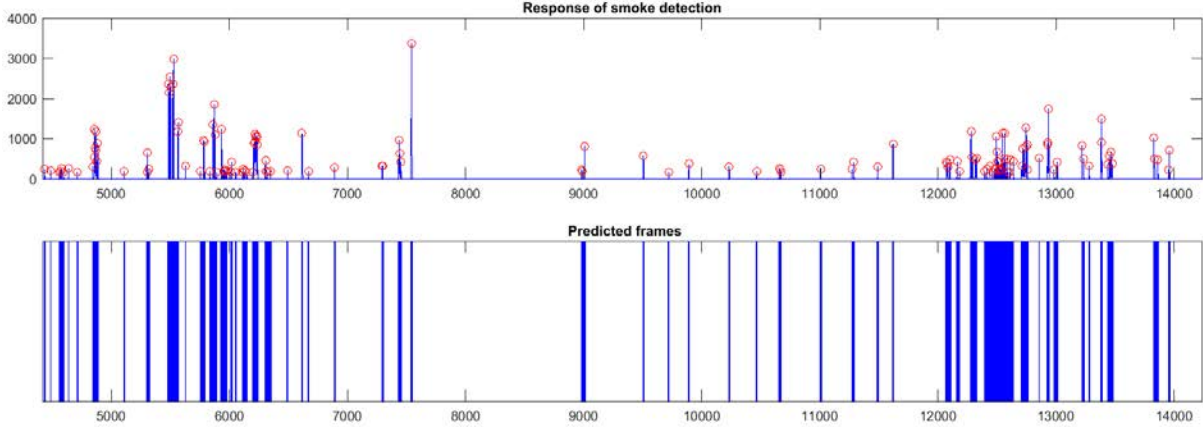


Figure A.6: Animated images generated by the time lapse viewer for May 2, 2015.



Table A1: Comparison of PM_{2.5} Levels at Avalon Monitor Before and After Camera.

Wind Direction (Degrees)	Before Camera	After Camera	Percent Change	t-stat (p-value)	Obs. Before/ Obs. After
205 - 215	13.793 ^A	11.193	18.85	2.568 (0.005)	116/135
200 - 220	13.862	12.649	8.75	1.937 (0.027)	443/387
195 - 225	13.904	13.038	6.23	1.595 (0.055)	636/551
175 - 245	14.340	13.293	7.30	3.122 (0.001)	1,446/1,305
165 - 255	14.184	13.234	6.70	3.419 (0.000)	1,990/1,771

A = before and after camera activation measurements of PM_{2.5} expressed in ug/m³.

Table A2: Full Regression Model Results for All Calls to ACHD.

Covariates	(1)	(2)	(3)	(4)
Cam. On	-0.0250 (0.201)	-0.0759 (0.224)	-0.0302 (0.226)	-0.0488 (0.175)
Cam. Active	0.462*** (0.0823)	0.233 (0.247)	-0.944 (1.409)	-0.827 (1.171)
Closure		-1.390*** (0.285)	-1.366*** (0.286)	-1.254*** (0.246)
Article		0.654 (0.519)	0.765 (0.540)	0.662*** (0.148)
Meeting		-0.532 (0.464)	-0.571 (0.469)	-0.629* (0.346)
PM2.5		0.0216 (0.100)	0.161 (0.158)	0.140 (0.135)
SO2		0.119** (0.0498)	0.0866 (0.0714)	0.0915 (0.0637)
O3		-0.0258 (0.135)	0.111 (0.217)	0.125 (0.175)
Incident		-0.0657 (0.203)	-0.0807 (0.200)	-0.0493 (0.150)
spring		-0.388** (0.166)	-0.358** (0.166)	-0.332** (0.142)
summer		-0.586*** (0.188)	-0.569*** (0.185)	-0.485*** (0.154)
fall		-0.320** (0.154)	-0.278* (0.155)	-0.213* (0.125)
Day of sample		0.00767*** (0.00226)	0.00595** (0.00266)	0.00545** (0.00236)
(Day of sample)²		-0.0000191** (0.00000835)	-0.0000122 (0.0000100)	-0.0000112 (0.00000863)
(Day of sample)³		1.53e-08** (7.59e-09)	8.94e-09 (9.12e-09)	8.01e-09 (7.78e-09)
Temperature		0.0236** (0.00919)	0.0222** (0.00922)	0.0227*** (0.00760)
Temperature²		-0.0000121 (0.000390)	0.0000273 (0.000387)	-0.000161 (0.000332)
Wind		0.00271 (0.00441)	0.000682 (0.00559)	0.00626 (0.00442)
Wind		-0.00000629 (0.0000109)	-0.00000279 (0.0000137)	-0.0000142 (0.0000108)
Direction²				

Wind		-0.0449	-0.0142	-0.0247
Speed		(0.0905)	(0.0904)	(0.0722)
Wind		0.00949	0.00674	0.00659
Speed²		(0.00772)	(0.00764)	(0.00619)
Camera x PM2.5			-0.237	-0.154
			(0.185)	(0.159)
Camera x SO2			0.0677	0.0582
			(0.0997)	(0.0853)
Camera x O3			-0.234	-0.271
			(0.275)	(0.226)
Camera x Wind			0.00601	-0.00151
Direction			(0.00812)	(0.00667)
Camera x Wind			-0.0000117	0.00000445
Direction²			(0.0000201)	(0.0000166)
Camera x Monday				0.310**
				(0.143)
Camera x Weekday				0.582***
				(0.0893)
Monday				1.648***
				(0.100)
Weekday				1.060***
				(0.0639)
Constant	1.009***	-0.419	0.00154	-1.421
	(0.0565)	(0.708)	(1.084)	(0.892)
Ln(alpha)	-0.526***	-0.698***	-0.710***	-1.429***
	(0.0652)	(0.0640)	(0.0634)	(0.104)
N	730	727	727	727

Standard errors in parentheses: * p<0.10, ** p<0.05, *** p<0.01

Dependent Variable is daily count of all calls to ACHD pertaining to air pollution.

Table A3: Full Regression Model Results for Shenango-Related Calls to ACHD.

Covariates	(1)	(2)	(3)	(4)
Cam. On	0.337 (0.313)	0.519 (0.432)	0.653 (0.445)	0.704* (0.392)
Cam. Active	0.280** (0.131)	-0.184 (0.331)	-4.700* (2.456)	-3.718 (2.491)
Closure		-1.992*** (0.562)	-1.821*** (0.577)	-1.809*** (0.553)
Article		0.498 (0.608)	0.623 (0.612)	0.250 (0.298)
Meeting		-0.916 (0.640)	-0.877 (0.634)	-1.012** (0.497)
PM2.5		0.134 (0.177)	0.0977 (0.267)	0.144 (0.239)
SO2		0.103 (0.0811)	0.0968 (0.118)	0.0985 (0.110)
O3		-0.176 (0.199)	0.237 (0.359)	0.313 (0.324)
Incident		0.0559 (0.334)	0.000337 (0.339)	0.0615 (0.295)
spring		-0.208 (0.274)	-0.131 (0.267)	-0.0610 (0.221)
summer		-0.596* (0.324)	-0.563* (0.319)	-0.437 (0.273)
fall		-0.0771 (0.254)	0.00192 (0.256)	0.0905 (0.211)
Day of sample		0.0102** (0.00401)	0.00523 (0.00517)	0.00452 (0.00512)
(Day of sample) ²		-0.0000217 (0.0000133)	-0.000000877 (0.0000184)	0.00000128 (0.0000178)
(Day of sample) ³		1.55e-08 (1.18e-08)	-4.08e-09 (1.67e-08)	-5.96e-09 (1.59e-08)
Temperature		0.0355** (0.0165)	0.0354** (0.0163)	0.0349*** (0.0131)
Temperature ²		-0.000478 (0.000676)	-0.000483 (0.000671)	-0.000603 (0.000585)
Wind		0.0220** (0.00897)	0.0182 (0.0117)	0.0301*** (0.0106)
Direction		-0.0000526** (0.0000220)	-0.0000453 (0.0000290)	-0.0000696*** (0.0000259)
Wind ²		-0.0212 (0.0000220)	0.0284 (0.0000290)	0.0213 (0.0000259)

Speed		(0.149)	(0.151)	(0.131)
Wind		0.0117	0.00726	0.00609
Speed²		(0.0125)	(0.0126)	(0.0110)
Camera x PM2.5			0.0682	0.142
			(0.305)	(0.290)
Camera x SO2			0.00928	0.0325
			(0.155)	(0.139)
Camera x O3			-0.761*	-1.038**
			(0.451)	(0.439)
Camera x Wind			0.0131	-0.00311
Direction			(0.0152)	(0.0137)
Camera x Wind			-0.0000274	0.00000766
Direction²			(0.0000376)	(0.0000338)
Camera x Monday				-0.706
				(0.893)
Camera x Weekday				-0.249
				(0.886)
Monday				5.099***
				(0.703)
Weekday				4.236***
				(0.697)
Constant	0.130	-4.605***	-2.631	-7.899***
	(0.0990)	(1.288)	(1.869)	(1.854)
Ln(alpha)	0.882***	0.666***	0.657***	-0.188
	(0.0905)	(0.0970)	(0.0967)	(0.131)
N	730	727	727	727

Standard errors in parentheses: * p<0.10, ** p<0.05, *** p<0.01

Dependent Variable is daily count of all calls to ACHD pertaining to Shenango Mill.

Table A4: Full Regression Model Results for Steel Manufacturing Calls to ACHD

Covariates	(1)	(2)	(3)	(4)
Cam. On	-1.096 (0.770)	0.744 (0.770)	1.088 (0.785)	1.116 (0.807)
Cam. Active	1.738*** (0.302)	-0.839 (0.957)	-1.513 (6.400)	-1.749 (5.814)
Closure		0.328 (0.761)	0.445 (0.786)	0.432 (0.726)
Article		0.675 (0.838)	0.831 (0.890)	0.923* (0.493)
Meeting		-0.123 (0.805)	-0.0389 (0.823)	-0.130 (0.797)
PM2.5		-0.206 (0.286)	-0.195 (0.686)	-0.217 (0.640)
SO2		0.185 (0.141)	0.550 (0.412)	0.540 (0.365)
O3		0.0111 (0.433)	1.132 (1.089)	1.177 (1.100)
Incident		-0.655 (0.582)	-0.688 (0.603)	-0.670 (0.581)
spring		2.220*** (0.635)	2.145*** (0.648)	2.280*** (0.682)
summer		1.853*** (0.632)	1.712*** (0.664)	1.811*** (0.668)
fall		0.503 (0.486)	0.407 (0.503)	0.521 (0.522)
Day of sample		0.0177 (0.0175)	0.0122 (0.0179)	0.0127 (0.0176)
(Day of sample) ²		-0.0000246 (0.0000495)	0.00000228 (0.0000565)	0.00000136 (0.0000542)
(Day of sample) ³		1.62e-08 (4.07e-08)	-9.47e-09 (4.93e-08)	-8.35e-09 (4.69e-08)
Temperature		0.0831* (0.0491)	0.0770 (0.0492)	0.0571 (0.0490)
Temperature ²		-0.00398** (0.00181)	-0.00396** (0.00177)	-0.00343* (0.00176)
Wind		-0.00427 (0.0128)	0.0314 (0.0244)	0.0429 (0.0263)
Wind		0.0000124 (0.0000312)	-0.0000856 (0.0000628)	-0.000110* (0.0000664)
Direction ²				

Wind		0.306	0.261	0.240
Speed		(0.456)	(0.398)	(0.347)
Wind		-0.0329	-0.0293	-0.0268
Speed²		(0.0460)	(0.0425)	(0.0351)
Camera x PM2.5			-0.0487	0.0286
			(0.702)	(0.673)
Camera x SO2			-0.427	-0.407
			(0.436)	(0.395)
Camera x O3			-1.307	-1.280
			(1.247)	(1.245)
Camera x Wind			-0.0373	-0.0509*
Direction			(0.0282)	(0.0297)
Camera x Wind			0.000104	0.000134*
Direction²			(0.0000720)	(0.0000749)
Camera x Monday				1.026*
				(0.536)
Camera x Weekday				1.530***
				(0.487)
Monday				16.18***
				(0.461)
Weekday				15.45***
				(0.343)
Constant	-3.064***	-7.361**	-7.408	-23.72***
	(0.265)	(3.285)	(5.815)	(5.327)
Ln(alpha)	1.025***	0.625*	0.517	-0.0739
	(0.301)	(0.349)	(0.350)	(0.420)
N	730	727	727	727

Standard errors in parentheses. * p<0.10, ** p<0.05, *** p<0.01

Dependent Variable is daily count of Steel Manufacturing calls to ACHD pertaining to air pollution.

Table A5: Comparison of Different Estimators for All Calls to ACHD.

Covariates	(1)	(2)	(3)
	OLS	Neg. Bin.	Poisson
Camera On	-0.146 (0.780)	-0.0488 (0.175)	-0.0571 (0.198)
Closure	-5.027*** (1.261)	-1.254*** (0.246)	-1.338*** (0.248)
Article	6.170*** (2.367)	0.662*** (0.148)	0.697*** (0.167)
Camera Active	-5.682 (5.436)	-0.827 (1.171)	-0.954 (1.332)
Camera x Monday	1.918* (0.979)	0.310** (0.143)	0.283* (0.151)
Camera x Weekday	2.448*** (0.405)	0.582*** (0.0893)	0.583*** (0.0944)
Monday	4.372*** (0.585)	1.648*** (0.100)	1.679*** (0.105)
Weekday	1.961*** (0.208)	1.060*** (0.0639)	1.084*** (0.0680)
Constant	-2.534 (3.496)	-1.421 (0.892)	-1.507 (1.058)
Season Fixed Effects	X	X	X
Cubic Time Trend	X	X	X
Weather Controls	X	X	X
Monitor Pollution	X	X	X
Camera x Monitor Pollution	X	X	X
Inalpha		-1.429*** (0.104)	
Adj. R²	0.306		
N	727	727	727

Standard errors in parentheses

Note: * p<0.10, ** p<0.05, *** p<0.01

Table A6: Comparison of Different Estimators for Shenango-Related Calls to ACHD.

Covariates	(1) OLS	(2) Neg. Bin.	(3) Poisson
Camera On	0.643 (0.610)	0.704* (0.392)	0.389 (0.325)
Closure	-1.907*** (0.595)	-1.809*** (0.553)	-1.899*** (0.504)
Article	1.884*** (0.439)	0.250 (0.298)	0.359 (0.264)
Camera Active	-5.420* (3.000)	-3.718 (2.491)	-2.340 (2.355)
Camera x Monday	0.133 (0.599)	-0.706 (0.893)	-0.766 (0.936)
Camera x Weekday	0.703*** (0.242)	-0.249 (0.886)	-0.293 (0.925)
Monday	2.807*** (0.451)	5.099*** (0.703)	5.075*** (0.713)
Weekday	1.276*** (0.151)	4.236*** (0.697)	4.223*** (0.704)
Constant	-3.040 (2.371)	-7.899*** (1.854)	-7.336*** (1.991)
Season Fixed Effects	X	X	X
Cubic Time Trend	X	X	X
Weather Controls	X	X	X
Monitor Pollution	X	X	X
Camera x Monitor Pollution	X	X	X
Ln(alpha)		-0.188 (0.131)	
Adj. R²	0.234		
N	727	727	727

Standard errors in parentheses

Note: * p<0.10, ** p<0.05, *** p<0.01

Table A7: Comparison of Different Estimators for Steel Manufacturing Calls to ACHD.

Covariates	(1) OLS	(2) Neg. Bin.	(3) Poisson
Camera On	0.0201 (0.0631)	1.116 (0.807)	1.103 (0.828)
Closure	0.207 (0.168)	0.432 (0.726)	0.422 (0.750)
Article	0.378 (0.256)	0.923* (0.493)	0.975** (0.488)
Camera Active	-0.366 (0.721)	-1.749 (5.814)	-1.710 (6.012)
Camera x Monday	0.274*** (0.0965)	1.026* (0.536)	1.205** (0.480)
Camera x Weekday	0.268*** (0.0594)	1.530*** (0.487)	1.687*** (0.373)
Monday	0.124*** (0.0449)	16.18*** (0.461)	15.26*** (0.452)
Weekday	0.0759*** (0.0220)	15.45*** (0.343)	14.59*** (0.356)
Constant	0.0897 (0.241)	-23.72*** (5.327)	-22.90*** (5.407)
Season Fixed Effects	X	X	X
Cubic Time Trend	X	X	X
Weather Controls	X	X	X
Monitor Pollution	X	X	X
Camera x Monitor Pollution	X	X	X
Ln(alpha)		-0.0739 (0.420)	
Adj. R²	0.100		
N	727	727	727

Standard errors in parentheses

Note: * p<0.10, ** p<0.05, *** p<0.01