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ABSTRACT

This paper analyzes the effects of environmental policy on employment (and unemployment) using a new general-equilibrium two-sector search model. We find that imposing a pollution tax causes substantial reductions in employment in the regulated (polluting) industry, but this is offset by increased employment in the unregulated (nonpolluting) sector. Thus the policy causes a substantial shift in employment between industries, but the net effect on overall employment (and unemployment) is small, even in the short run. An environmental performance standard causes a substantially smaller sectoral shift in employment than the emissions tax, with roughly similar net effects. The effects on the unregulated industry suggest that empirical studies of environmental regulation that focus only on regulated firms can be misleading (and those that use nonregulated firms as controls for regulated firms will be even more misleading). The paper's results also suggest that overall effects on employment are not a major issue for environmental policy, and that policymakers who want to minimize sectoral shifts in employment might prefer performance standards over environmental taxes.

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1. Introduction

Effects on employment have played a central role in the political debate over environmental regulations, especially during the recent economic downturn, with opponents deriding regulations as "job killers" and proponents touting "green jobs." This focus is understandable, given the large potential welfare effects of involuntary unemployment. However, economic studies of the effects of environmental regulation do not adequately answer the question of how regulation affects unemployment.

A substantial number of empirical studies have looked at how regulations affect employment.¹ But while such studies provide valuable information on employment changes in regulated industries, they do not measure the effects on unregulated industries, so to the extent that regulation affects employment in those other industries, such studies cannot measure the overall effect. A more serious problem is that these studies often employ a difference-indifferences approach, using firms in unregulated industries (or firms in regulated industries, but in areas that are unregulated or subject to less stringent regulation) as controls. To the extent that regulation affects employment at those firms, such studies will not only miss the effects on unregulated firms, but also yield biased estimates of the effects on regulated firms.

Addressing those issues requires a general-equilibrium analysis. But existing generalequilibrium models used to analyze environmental regulation almost always assume full employment. And the few models in this area that do allow for unemployment typically focus on types of unemployment that are largely unimportant in the United States (e.g., unemployment caused by strong unions that negotiate wages well above free-market levels).²

This paper develops a new model to study how environmental regulation affects employment and unemployment. It incorporates a search model with frictions as in Mortensen and Pissarides (1994), together with a simple two-sector general-equilibrium model of environmental policy, roughly calibrated to correspond to the effects of imposing a carbon policy in the United States. In this model, unemployed workers must search and match with job openings in each period; the interaction of carbon policy with this matching process plays a key role in determining the economy-wide employment impacts of the regulation.

¹ For examples, see Berman and Bui (2001), Curtis (2012), Greenstone (2002), and Morgenstern et al. (2002)

² See, for example, Bovenberg and van der Ploeg (1996).

This paper makes three substantial contributions. First, we show that while imposing a pollution tax leads to substantial reductions in employment in the polluting sector of the economy, those losses are offset by an employment increase of similar magnitude in the nonpolluting sector, driven both by consumer substitution from polluting to nonpolluting goods and by decreased competition for workers from the polluting sector, which makes it easier for the nonpolluting sector to hire. Consequently, while there is a substantial shift in employment between industries, the net effect on unemployment is small, even in the short run. Empirical studies that look only at regulated firms would greatly overstate that net effect (though they can measure the job loss in the regulated sector).³ Difference-in-differences studies that use unregulated firms as a control group would be even further off, not only missing job gains in unregulated firms, but also seriously overstating job losses in regulated firms. Our results suggest that those studies could overstate effects on regulated firms by a factor of almost two and overstate the overall net effects by far more. This emphasizes the importance of considering general-equilibrium effects.

Second, we show that the magnitude of the employment shift (both the job losses in the polluting sector and gains in the nonpolluting sector) is much smaller under a performance standard (a constraint on pollution emissions per unit of output) than under a pollution tax. A performance standard is equivalent to a tax on emissions and subsidy on output in the dirty industry (see, for example, Holland et al. 2009; Fullerton and Metcalf 2001). As a result, the price increase for polluting goods is much smaller under a performance standard than under an equivalent emissions tax, and thus the substitution in consumption and corresponding shift in employment is correspondingly smaller. This suggests that, to the extent that policymakers want to minimize sectoral shifts in employment, performance standards (and related intensity-standard policies) may be attractive.

Third, the paper develops and demonstrates a tractable framework for bringing unemployment into computable general equilibrium (CGE) models of environmental regulation.⁴

³ This practice—interpreting estimated effects on the regulated sector as representing the overall effect on employment—is relatively common. For example, several recent EPA regulatory impact analyses estimate effects on overall employment based on coefficients from Morgenstern et al. (2002) for the effects on regulated industries. Smith et al. (2013) provides an extensive discussion of how EPA has estimated employment effects in regulatory impact analyses.

⁴ Balistreri (2002) provides an alternative approach for integrating unemployment into a CGE model. That paper uses a reduced-form representation, rather than explicitly modeling search as we do. Our approach has a number of

This general framework could be useful for studying a wide range of other questions involving unemployment and environmental regulations, such as questions about the optimal timing of regulations (e.g., should a major new environmental regulation be implemented during recession, or would it be better to wait until after the economy has rebounded?), about the design of regulations (e.g., how is the choice of price-based versus quantity-based regulations affected by unemployment concerns?), or about the distributional implications of changes in employment patterns resulting from regulation. And it could readily be adapted to look at employment effects of other types of policies, such as trade policy, where employment effects play an important political role.

The rest of the paper is organized as follows. Section 2 presents the structure of the model. Section 3 explains the calibration of the model, and Section 4 discusses the policy simulations and results. Section 5 considers extensions of the baseline model, and the final section offers conclusions.

2. A Two-Sector Model

We begin by introducing a relatively simple two-sector model that will enable us to analyze the key channels through which environmental policies affect employment in both the regulated and unregulated sectors. The key element is the inclusion of a search friction as in Pissarides (1985) and Mortensen and Pissarides (1994). The model is a two-sector extension of the Shimer (2010) search model with variable hours.⁵ The two sectors are the clean sector, *c*, and the dirty sector, *d*.

advantages, including the ability to look at temporary effects on unemployment resulting from sectoral shifts in demand. Shimer (2013) also looks at employment, with a focus on sectoral shifts caused by environmental policy, but again does not explicitly model search (instead, workers must spend an exogenous amount of time unemployed any time they switch industries) and focuses on a different question (whether labor transition costs should affect the optimal emissions tax). Aubert and Chiroleu-Assouline (2015) explicitly model employment search in a model of environmental taxation but focus only on the long-run steady state, rather than looking at transitions.

⁵ To our knowledge, Yedid-Levi (2013) is the only other paper to extend Shimer's (2010) search model to multiple sectors. It introduces a two-sector model with search and matching frictions with output vacancy costs, wage rigidities, and capital adjustment costs. Yedid-Levi focuses on a very different problem (explaining sector co-movement) than we do and thus models a different pair of industries (consumption and investment goods, as opposed to polluting and nonpolluting goods). Our model also differs in a number of other important respects. First, we use the labor vacancy cost as in Shimer (2010). Second, we exclude Yedid-Levi's characterization of wage rigidities—in his model, wage rigidities are introduced through pro-cyclical bargaining power, but our model lacks cyclicality—and we also exclude capital entirely. Third, we formulate the sector-specific job-matching functions differently than did Yedid-Levi (see the discussion in section 2.1, below).

Hiring is subject to a matching process between recruiters and unemployed workers. There are no on-the-job searches and no job-to-job transitions: only unemployed workers can search for jobs. Unemployed workers can search for a job in both sectors. An exogenous and constant fraction of job matches dissolves each period. All workers are members of a representative household that provides perfect insurance to its members. This representative household owns the firms.

2.1. Matching Process

Without loss of generality, we normalize the measure of workers to one. Let n_j denote the measure of workers in sector $j = \{c, d\}$. Aggregate employment is given by $\overline{n} = n_c + n_d$; the unemployment rate (and the measure of unemployed workers) is given by $1 - \overline{n}$. Unemployed workers search indiscriminately across sectors. Let v_j denote the number of recruiters in each sector. One recruiter, working h_j hours, can hire $H_j(\overline{\theta})$ unemployed workers per hour. Labor market tightness, $\overline{\theta}$, is the ratio of aggregate recruiting effort to the number of unemployed workers, $\overline{\theta} = \sum_j (v_j h_j)/(1-\overline{n})$. Recruiting productivity is therefore endogenous and is a function of total recruiting activity and the number of unemployed workers. To define recruiting productivity, we start by defining the matching function. The number of matches in each sector, m_j , is function of recruiting effort in each sector, $v_j h_j$, aggregate recruiting effort, and the number of people looking for jobs. We follow Shimer (2010) and assume a constant-returns-to-scale matching function but extend that function to accommodate multiple sectors such that

$$m_j = \mu_j (1 - \overline{n})^{\gamma_j} (v_j h_j) \left(\sum_i v_i h_i \right)^{-\gamma_j}$$
(1)

where μ_j and γ_j are the matching efficiency and matching elasticity parameters for each sector.⁶ Matches in a given sector are increasing in the number of workers searching for jobs and the sector's own recruiting effort. Matches in one sector are decreasing in the recruiting effort of the other sector as more competition for workers reduces the probability of a match. This represents

⁶Note that this matching function specification reduces to a Cobb-Douglas matching function (which is standard in the job-matching literature) when there is only one sector.

an important departure from Yedid-Levi (2013), which assumes that matches in one sector are independent of recruiting effort in the other sector. That assumption seems unrealistic, given that both sectors are hiring from the same pool of workers.

The recruiting technology implies that the number of matches in a sector must be equal to recruiting effort times the recruiting productivity, $m_j = (v_j h_j)H_j$, and the number of matches in a sector must equal the number of workers searching times the probability a worker finds a job in that sector, $m_j = (1 - \overline{n})\phi_j$. Using equation (1), we can define sector recruiting productivity, H_j , and the probability of finding a job in sector *j*, ϕ_i , as

$$H_{i} = \mu_{i} (\overline{\theta})^{-\gamma_{j}} \tag{2}$$

$$\phi_{i} = \mu_{i} \theta_{i} \overline{\theta}^{-\gamma_{j}} \tag{3}$$

where $\theta_i = (v_i h_i)/(1-\overline{n})$ is the ratio of sector-specific recruiting effort to unemployed workers.

2.2. Households

We use a representative household framework and assume full insurance within the household. This assumption implies that the marginal utility of consumption is equalized across workers and is independent of both past and current employment status.⁷ Employed workers in sector *j* receive an hourly wage w_j and work h_j hours. Given labor income taxes at the rate τ_L , after-tax earnings per worker in sector *j* are given by $(1-\tau_L)w_jh_j$. Unemployed workers work zero hours and receive unemployment compensation *b*, which is held constant in real terms over time. The household owns the firms and has access to a complete set of state-contingent claims. Let *Q* denote the price of an Arrow security that delivers one unit of consumption the following period.

Members of the household gain utility from consumption and disutility from work. The period felicity function is

$$U(c,h) = \log c - \psi \frac{h^{1+\chi}}{1+\chi}$$
(4)

⁷ This common assumption within the search literature, due originally to Merz (1995), greatly simplifies the problem of the household. Blundell et al. (2008) find evidence that only poor households cannot fully insure against transitory shocks.

where ψ represents a disutility from work parameter and $1/\chi$ is the Fritsch elasticity of labor supply.⁸ Given that preferences are separable in consumption and hours worked, consumption will be independent of the employment status of the worker.

The household discounts future utility with a discount factor β . Households take the job finding rates ϕ_j as given. The problem of the household is to maximize lifetime-discounted utility. This problem can be written dynamically as a Bellman equation, where the state of the household is given by the value of its current assets, *B*, and the distribution of workers across sectors, n_j .

$$V(B, n_c, n_d) = \max_{c, B'} \{ n_c U(c, h_c) + n_d U(c, h_d) + (1 - \overline{n}) U(c, 0) + \beta E[V(B', n_c', n_d')] \}$$
(5)

subject to the budget constraint,

$$P^{C}c + QB' \le (1 - \tau_{L})n_{c}w_{c}h_{c} + (1 - \tau_{L})n_{d}w_{d}h_{d} + (1 - \overline{n})P^{C}b + B - T$$
(6)

and the law of motion of employment

$$n'_{j} = (1 - \pi_{j})n_{j} + \phi_{j}(1 - \overline{n}), \quad j = c, d$$
(7)

where P^{C} is the price of the consumption good, *T* is government lump-sum taxes, and π_{j} is the exogenous rate of job destruction in each sector.

Let λ denote the Lagrange multiplier on the budget constraint. By combining the firstorder condition with respect to next period's assets with the envelope condition with respect to current assets, the Euler equation can be written as

$$Q = \beta \frac{\lambda'}{\lambda} \tag{8}$$

The marginal value of having a worker employed in a given sector is given by the envelope condition with respect to the number of workers employed in that sector,

$$V_{n_j} = U(c,h_j) - U(c,0) + \lambda [(1 - \tau_L)w_j h_j - b] + (1 - \pi_j)\beta E V'_{n_j} - \overline{\phi}$$
(9)

where $\overline{\phi} = \phi_c \beta E V'_{n_c} + \phi_b \beta E V'_{n_d}$. The marginal value consists of utility and compensation differentials between employed workers in sector *j* and unemployed workers. The value also includes the continuation value of being employed in the same job the following period. The

⁸ Note that we are assuming that workers get the same disutility from work in both sectors.

term $\overline{\phi}$ represents the opportunity cost of being employed (i.e., being unable to look for other jobs).

2.3. Firms

Firms in each sector produce a sector-specific good using only a labor input, $y_j = A_j h_j l_j$, where l_j denotes the number of workers using a production technology that produces A_j units of output per hour worked. Given a measure of total workers n_j in the beginning of the period, the firm must decide to allocate workers to production, l_j , or recruitment, v_j , such that $l_j + v_j = n_j$.⁹ Pollution emissions from production are given by

$$e_j = (1 - v_j)\mu_j^e \overline{y}_j \tag{10}$$

where unconstrained emissions are a fixed multiple μ_j^e of net output, \overline{y}_j , and v_j is the fraction of unconstrained emissions that are abated. Net output is equal to gross output minus abatement costs, $\overline{y}_j = y_j - z_j$. The cost of abatement is modeled as a per (gross) unit of output cost \overline{z} such that

$$z_j = \overline{z}(v_j) y_j. \tag{11}$$

The domain of the abatement function \overline{z} is [0,1]. We assume that $\overline{z}(0) = 0$, $\lim_{v \to 1} \overline{z}(v) = \infty$, $\overline{z}'(0) = 0$ and $\overline{z}''(v) > 0$. Firms must pay an emissions tax τ^e for all emissions that are not abated.¹⁰

Firms are owned by households and discount future-period profits with the discount factor Q from the household problem. Following Shimer (2010), we define the recruiter ratio $\overline{v}_j = v_j / n_j$. Net output can be written as a function of total workers, the recruiter ratio, and abatement, $y_j = A_j h_j n_j (1 - \overline{v}_j)(1 - \overline{z}(v_j))$. Firms take the endogenous recruiting productivity, H_j , as given when making recruiter decisions. The dynamic problem of the firm is to choose the recruiter ratio and abatement to maximize the value of the firm. The Bellman equation is

⁹ Workers receive the same wage and work the same hours regardless of position. Therefore, workers are indifferent between job types.

¹⁰ If the emissions tax is zero (or if firms have no emissions, as is the case in the clean sector), firms will optimally set abatement to be zero.

$$J(n_{j}) = \max_{\overline{v}_{j} \ge 0, v_{j} \ge 0} \{ p_{j}^{n}(v_{j}) A_{j} h_{j} n_{j} (1 - \overline{v}_{j}) (1 - \overline{z}(v_{j})) - (1 + \tau_{p}) n_{j} h_{j} w_{j} + \mathbb{E}[QJ(n_{j}')] \}$$
(11)

subject to

$$n'_{j} = (1 - \pi_{j})n_{j} + H_{j}\overline{v}_{j}h_{j}n_{j}$$

$$\tag{12}$$

where $p_j^n(v_j)$ denotes the net price received by sector *j* for its output and τ_p denotes the payroll tax. The net price of output is equal to the price of the good minus the emissions tax payments, $p_j^n(v_j) = p_j - \tau^e \mu_j^e (1 - v_j)$. The first-order condition with respect to the recruiter ratio sets the number of recruiters such that the marginal cost of switching an additional worker from production to recruitment equals the expected value of recruitment:

$$p_j^n(\mathbf{v}_j)A_j(1-\overline{z}(\mathbf{v}_j)) = H_j \mathbb{E}[QJ'_{n_i}]$$
(13)

where J'_{n_j} is the value (to the firm) of an additional worker in the following period. The value of an additional worker can be derived from the envelope condition with respect to the number of workers and substituting out equation (14):

$$J_{n_j} = p_j^n(v_j) A_j h_j (1 - \overline{z}(v_j)) - (1 + \tau_P) w_j h_j + (1 - \pi_j) \mathbb{E}[QJ'_{n_j}].$$
(14)

This marginal value of a worker is equal to the marginal revenue minus compensation plus the expected continuation value times the probability that the match does not exogenously dissolve at the end of the period. The first-order condition with respect to abatement will equate the marginal cost of abatement to the marginal benefit of abatement (i.e., lower emissions tax payments) such that

$$p_j^n(\mathbf{v}_j)\overline{z}'(\mathbf{v}_j) = \tau^e \mu^e (1 - \overline{z}(\mathbf{v}_j)).$$
⁽¹⁵⁾

2.4. Wage Bargaining

In each period, the firm and its workers enter a bargaining process to determine hours and the hourly wage. Following Shimer (2010), we assume a Nash bargaining process.¹¹ Let η denote the bargaining power of the employer. The equilibrium wages and hours are the solution to

$$\max_{w_j,h_j} J^{\eta}_{n_j} V^{1-\eta}_{n_j} \tag{16}$$

¹¹ The derivation of the wage and hours equations follows exactly from section 2.4.2 in Shimer (2010). Adding a second sector does not significantly change the derivation, and therefore we omit the derivation here for brevity.

Nash bargaining implies that hours per worker are set to maximize the value of the match surplus. This means that the marginal value of an additional hour of work is equal to the disutility of an additional hour of work,

$$(1+\tau_P)\psi h_j^{\chi} = (1+\tau_L)\lambda p^n(v_j)A_j(1-\overline{z}(v_j))$$
(17)

Nash bargaining also implies that the match surplus is divided according to a constant share rule. The equilibrium after-tax pay packet for a worker is

$$(1-\tau_L)h_jw_j = \frac{1-\tau_L}{1+\tau_P}(1-\eta) \left[p_j^n(v_j)A_jh_j(1-\overline{z}(v_j)) + \left[\sum_i p_i^n A_i(1-\overline{z}(v_i))\theta_i \right] \right] + \eta \left[\frac{\psi}{\lambda} \frac{h^{1+\chi}}{1+\chi} + P^c b \right].$$
(18)

The first term on the right-hand side of the equation reflects the marginal revenue of an additional employee, $p_j^n(v_j)A_jh_j(1-\overline{z}(v_j))$, and the ease with which job searchers can find jobs, $\sum_i p_i^n A_i(1-\overline{z}(v_i))\theta_i$. Workers need to be compensated for their opportunity cost of not searching for other jobs, and this compensation is increasing in the ease at which workers can find other jobs in either sector. As a result, a decrease in recruitment in one sector will decrease the equilibrium wages in the other sector because the bargaining strength of workers in the other sector has been reduced. The second term is the flow value of unemployment: workers need to be compensated for the loss of unemployment benefits as well as the differences in utility that would occur if an unemployed worker took the job as opposed to remaining unemployed.

2.5. Government

We assume that the government runs a balanced budget such that the total value of labor, payroll, and lump-sum taxes exactly equals the value of unemployment benefits paid out each period,

$$(\tau_L + \tau_P)(n_c w_c h_c + n_d w_d h_d) + \tau^e(e_d + e_c) + T = (1 - \overline{n})b$$
(18)

where $\tau^{e}(e_{d} + e_{c})$ represents revenues from the tax on unabated emissions.

2.6. Emissions Taxes and Performance Standards

For simplicity, we assume that the clean sector has zero emissions and therefore $v_c = 0$

regardless of policy. When τ^e equals zero (the business-as-usual case), the dirty sector has zero incentive to reduce its emissions and therefore abatement is zero ($v_d = 0$). When τ^e is positive, the dirty sector will reduce emissions until the marginal cost of abatement equals the marginal benefit of reducing the emissions tax (as seen in equation (16)). As we always assume a balanced budget, the introduction of an emissions tax (or an increase in the emissions tax) will require decreases in lump-sum taxes, labor taxes, payroll taxes, or some combination of all three taxes.

Performance standards are policies that restrict the dirty sector's emissions per unit of output. Rewriting equation (10), emissions per unit output are simply equal to the fixed emissions coefficient and abatement, $e_j / y_j = (1 - v_j)\mu_j^e$. A performance standard is modeled as a restriction on emissions per unit of output such that they must be less than or equal to some maximum, $e_j / y_j \le \overline{e_j}$. This constraint is isomorphic to a minimum constraint on abatement and can be implemented directly using a constrained maximization problem; equation (16) is replaced by

$$p_j A_j (1 - \overline{v}_j) h_j n_z \overline{z'}(v_j) = \lambda_j^e \mu_j^e$$
⁽¹⁹⁾

where λ_j^e is the shadow value of the constraint.

2.7. Consumption Good and Market Clearing

The consumption good is created by aggregating the output of the two sectors according to a constant elasticity of substitution (CES) aggregation function. The price of the consumption good is

$$P^{c} = (\gamma^{c})^{\frac{\sigma^{c}}{1-\sigma^{c}}} \left[\sum_{j} \alpha_{j}^{\sigma^{c}} p_{j}^{1-\sigma^{c}} \right]^{\frac{1}{1-\sigma^{c}}}$$
(20)

where α_j is the CES share parameter for good *j*, γ^c is the CES scaling parameter, and σ^c is the elasticity of substitution between the clean and dirty sectors in consumption. The amount of consumption demanded by each sector can then be written as

$$c_{j} = (\gamma^{c} \alpha_{j})^{\sigma^{c}} \left[\frac{p_{j}}{P^{c}} \right]^{-\sigma^{c}} C$$
(21)

Markets clear in each sector when net output is equal to consumption demand,

$$y_j = c_j \,. \tag{22}$$

3. Calibration

Parameter values are primarily taken from existing sources. The benchmark parameters of the model are summarized in Table 1. Our benchmark calibration assumes symmetry between the sectors for all values except for emissions rate and relative size.

3.1. Household

The length of a period is one month, to reflect the relatively rapid job turnover that occurs in the United States. The discount factor is $\beta = 0.996$, reflecting an average annualized rate of return of 4 percent. The Frisch elasticity of labor supply is $\chi = 1$. As discussed in Hall and Milgrom (2008) and Yedid-Levi (2013), this value represents an approximate average between the elasticity found for middle-age men (0.7) and higher elasticities found for women and young men. The disutility parameter ψ is set such that average time worked in the non-emissions-tax steady state is 0.33 (i.e., one-third of the time endowment). We set the tax on labor τ_L to be 0.25, which is approximately the average marginal income tax rate on labor income from federal and state income taxes combined.

We do not calibrate the value of unemployment benefits b to an empirical measure of the replacement rate. Rather, we find the value of b to satisfy the wage bargaining equation given the calibration of the other parameters. The implied value of b implies a replacement rate of approximately 42 percent (unemployment benefits as a percentage of after-tax earnings). This value is much larger than the 25 percent value used by Hall and Milgrom (2008), but it closely matches values used in other studies (for example, Amaral and Tasci 2013).

We assume an unemployment rate of 7 percent in the non-emissions-tax equilibrium. In 2012, personal consumption expenditures in the United States were approximately \$11.1 trillion¹²; we set consumption equal to \$11.1 trillion and then scale by \$11.1 trillion such that the unit of consumption in our non-emissions-tax steady state is one. Through our CES aggregation function, we assume that the clean and dirty outputs are imperfect substitutes in consumption.

¹² Source: BEA National Income and Product Accounts Tables; Table 1.1.5, http://bea.gov/iTable/index_nipa.cfm

We set the elasticity of substitution in our base case to be a conservative 0.75.¹³ We then vary that value in the sensitivity analysis.

Household		
Discount factor	eta	0.996
Frisch elasticity of labor supply	χ	1
Disutility of work	ψ	6.5133
Taxes on labor income	$ au_{\scriptscriptstyle L}$	0.25
Elasticity of subs. in consumption	$\boldsymbol{\sigma}^{c}$	0.75
Unemployment benefits	b	0.3054
Firms		
Productivity	А	3.2422
Payroll tax	$ au_{\scriptscriptstyle P}$	0.12
Exogenous separation rate	π	0.042
Abatement cost level parameter	a_1	1
Abatement cost curvature	a_2	2.8
Clean share of economy		0.8
Emissions rate	μ^{e}	0.0025
Matching		
Match elasticity	γ	0.5
Match efficiency	μ	3.7342
Bargaining power	η	0.5

Table 1. Parameter Values

3.2. Firms

We begin by assuming that the clean sector is 80 percent of both consumption and labor (this implies symmetric production productivity parameters).¹⁴ Given this assumption, we can then

¹³ The CES parameters $\alpha_c, \alpha_d, \gamma^c$ are calibrated conditional on the CES elasticity and consumption shares in the standard manner.

¹⁴ This may represent an overestimate of the fraction of the economy that would be regulated by a carbon tax. The fossil fuel extraction, electric power, natural gas distribution, and petroleum refining industries make up about 5 percent of US output. However, adding transportation and construction to that list brings the total to over 20 percent of output. To the extent that we overestimate the size of the dirty industry, our results can be interpreted as an upper bound on the aggregate employment effects.

derive total output and labor employed by sector. Using the steady state value of the recruiter share (derived in the following section), we calibrate productivity in each sector such that $A_j = c_j / (h_j n_j (1 - \overline{v}_j))$.

According to the Job Openings and Labor Turnover Survey (JOLTS), the average prerecession separation rate was 4.2 percent per month. We assume that both sectors have the same separation rate. We set the payroll tax τ_P to be 0.12.

In the United States in 2012, total energy-related CO₂ emissions were approximately 5.2 billion short tons¹⁵, implying approximately 0.0005 short tons of CO₂ emissions per dollar of consumption. Given that the dirty sector is the only source of emissions in our model and it represents 20 percent of consumption, the emissions factor for the dirty industry is $\mu^e = 0.0025$.

We use the abatement cost function of Nordhaus (2008), $z(v) = a_1 v^{a_2}$. The curvature parameter a_2 is set to 2.8 as in Nordhaus (2008) and Heutel (2012).¹⁶ Given the unit scale of the labor force, we simply set a_1 to be equal to one.¹⁷

3.3. Matching

Estimates vary for the matching elasticity parameter γ . Petrongolo and Pissarides (2001) estimated the value to be between 0.3 and 0.5. Later studies found a value near 0.75 (Hall 2005; Shimer 2005). As in Hall and Milgrom (2008), we choose a compromise value of 0.5 and set the value to be equal for both sectors. To ensure the Hosios condition in our benchmark calibration, we follow standard search models by setting the bargaining power equal to the match elasticity such that $\eta = 0.5$.

To calibrate μ_j we follow the method of Shimer (2010). First, we assume, based on Silva and Toledo (2009), that the cost of recruiting a worker is approximately 4 percent of one worker's

¹⁵ Source: U.S. Energy Information Administration, http://www.eia.gov/totalenergy/data/monthly/pdf/sec12_3.pdf. ¹⁶ We use this functional form because it is common in the literature. However, note that it violates our earlier

assumption that $\lim_{v \to 1} \overline{z}(v) = \infty$ (i.e., that the cost of abatement goes to infinity as emissions net of abatement goes to

zero). That would cause problems at high levels of the emissions tax rate (for a sufficiently high tax rate, firms would set abatement high enough that net emissions would be negative). But the function generates reasonable results for the rates we consider in our simulations, which are well below the level at which net emissions would be negative.

¹⁷ This implies that the dirty sector can reduce emissions by 19.31 percent at the expense of 1 percent of its gross output.

quarterly wage. Assuming that cost comes entirely in the form of recruiters' time implies that H_j equals 25 in the initial steady state. To keep employment constant in the steady state, one can easily show that $\overline{v}_j = \pi_j / (H_j h_j)$ in the steady state; approximately 0.5 percent of workers are engaged in recruiting in each sector in the non-emissions-tax steady state. Given the recruiter ratio, we can derive total recruiting effort for each sector, $\overline{v}_j h_j n_j$, and then derive aggregate market tightness $\overline{\theta}$. Using the definition for recruiting productivity, equation (2), match efficiency is calculated as $\mu_i = H_i \overline{\theta}^{\gamma}$.

4. Emissions Tax

We now use the model to simulate the effect of three different policy scenarios on unemployment, employment by sector, hiring, earnings and other labor market outcomes. The first two policy scenarios each include a 20/ton tax on CO₂ emissions. The first scenario uses lump-sum rebates, and the second reduces payroll taxes to keep the government budget balanced. The third scenario uses an environmental performance standard (a constraint on emissions per unit of output), set at a level that gives the same emissions reductions in the steady state as the first scenario (20/ton tax with lump-sum rebates). In all three scenarios, we hold the real level of unemployment benefits constant. Because the performance standard slightly discourages overall economic activity, thus reducing tax revenues, we increase labor taxes to balance the government budget in this case.

Each of these policies reduces steady state emissions by approximately 13.6 percent.¹⁸ Emissions taxes give the dirty sector an incentive to reduce emissions through both output reductions and emissions-intensity improvements. In contrast, because the performance standard is a constraint on emissions per unit of output, it provides no direct incentive to reduce output from the polluting sector, and thus emissions reductions under the performance standard come almost entirely from lower emissions intensity.¹⁹

Figure 1 displays the percentage of emissions reductions that can be attributed to output reductions. Carbon taxes use both channels of emissions reductions, with 21–24 percent of

¹⁸ The emissions reductions are slightly less in the payroll tax reduction emissions tax policy because the tax cut stimulates economic activity.

¹⁹ See Goulder et al. (2016) for further discussion of this difference in incentives between taxes and intensity standards and its economic implications.

emissions reductions due to output reductions in the dirty sector. Roughly 2 percent of emissions reductions in the performance standard policy come from output reductions. Because the carbon taxes equalize marginal costs across both channels for reducing emissions, they have lower direct costs of reducing emissions than the performance standard. However, as is shown in the next section, the much larger dirty-sector output reductions under the carbon tax imply a large labor-force shift between sectors.



Figure 1. Percentage of Emissions Reductions due to Output Reductions

4.1. Aggregate Unemployment and Employment by Sector

Figure 2 displays the effects of all three policies on aggregate unemployment. The carbon tax with lump-sum rebates causes the unemployment rate to increase by about 0.26 percentage points. Neither the carbon tax with payroll tax reductions nor the performance standard has any meaningful impact on the aggregate unemployment rate; the policies increase the unemployment rate by about 0.02 and 0.04 percentage points, respectively.

Figures 3a–3c show the change in the level of employment in the clean and the dirty sectors for all three policies. In both carbon tax policies, long-run employment falls substantially in the dirty sector but rises substantially in the clean sector. The carbon tax raises the price of the polluting good, thus inducing consumers to substitute toward the clean good. This causes the dirty industry to contract and the clean industry to expand. When carbon tax revenues are used to

cut payroll taxes, the payroll tax cut encourages employment in both sectors, making the drop in dirty-sector employment slightly smaller and the rise in clean-sector employment larger. In this case, the net effect on employment is very small.²⁰



Figure 2. Aggregate Unemployment Rate

Performance standards have a much smaller effect on the price of the dirty good than the carbon tax: both taxes and standards raise production costs for the dirty good (because some workers are used for emissions abatement), but the performance standard does not tax the remaining emissions and thus raises the price of the dirty good by much less. This much smaller price increase for the dirty good leads to a much smaller drop in quantity demanded of the polluting good. Moreover, the need for emissions abatement implies lower output per worker (an effect which occurs under all three policies, but which is larger for the performance standard, because it relies much more on reductions in the emissions/output ratio), so any given level of output requires more workers. If demand for the dirty good is relatively inelastic (as it is in our central-case estimates), this latter effect dominates, and thus employment actually rises slightly in the polluting industry. Again, though, the effect on clean-sector employment has the opposite sign (in this case, because consumers have less to spend on the clean good, and clean-good

²⁰ If carbon tax revenue is used to cut taxes on labor income, the results are identical to the case with payroll tax cuts because of the assumption that wages are flexible and renegotiated each period. If wages were fixed or slow to adjust, then the effect of a cut in labor taxes paid by employers would differ from that for cutting a tax paid by employees.

producers face more competition for workers).²¹



Figure 3: Employment Changes by Sector









(c) Performance Standard

²¹ If dirty-good demand were more elastic, the effects of the performance standard on clean- and dirty-sector employment would reverse. The drop in quantity demanded would dominate the rise in workers per unit of output, and dirty-sector employment would fall. In that case, clean-sector employment would rise (spending on the dirty good would fall, leaving more to spend on the clean good, and clean-sector employers would face less competition for workers).

The net effects depend largely on how the policy shift affects the value of a job match, which (holding parameters constant) depends on how the real after-tax wage changes. Under the carbon tax with payroll tax cuts, the carbon tax reduces the real wage (by raising the price of the dirty good), but the payroll tax cuts offset that effect, leaving the after-tax real wage nearly unchanged. Under the carbon tax with lump-sum rebates, that offsetting effect is missing, so the real wage falls significantly. Under the performance standard, there also is no payroll tax cut (indeed, the payroll tax increases very slightly), but the price increase for the dirty good is much smaller, leading to a relatively small drop in the real wage.²²

There are two key points to take away here. First, under all three policy scenarios, the employment effects represent much more of a shift in employment—fewer jobs in one sector but more jobs in the other—rather than a change in overall employment. This is particularly true for the carbon tax with payroll tax cuts and the performance standard, where the net effect on employment is very small. Second, the employment shifts are much smaller under the performance standard than under either carbon tax policy, because the relative output price effects are much smaller under the performance standard. To the extent that policymakers view industry-level employment contractions as undesirable—even if they are offset by employment gains elsewhere—this will make performance standards relatively attractive compared with carbon taxes.

These figures also illustrate the major problem noted earlier with empirical studies that use a difference-in-differences estimator, with unregulated (i.e., clean) industries as a control group. In our model, neither carbon taxes nor performance standards directly affect the clean industry, and yet every policy we consider causes changes in clean-industry employment that are similar in magnitude to the dirty-industry employment change. A difference-in-differences estimator would take the difference between the employment changes across the two sectors as an estimate of the effect of environmental policy. Thus in the tax cases, it would substantially overestimate the job loss in the dirty sector and overestimate the total job loss by even more.²³ In the performance-standard case, it would find a significant job gain, when the true total effect is a

²² These effects mirror similar results for taxes versus performance standards from the literature on tax interactions (in models with full employment). See, for example, Goulder et al. (1999) or Fullerton and Metcalf (2001).

²³ While our model focuses on clean and dirty industries, the same problem would arise for difference-in-differences estimators more generally whenever the two regulated and unregulated firms' products are substitutes and labor can move between regulated and unregulated firms.

job loss.

4.2. Labor Market Outcomes

Table 2 displays the impact of the three policies on key labor-market variables in both the initial period of the policy (Pd 0) and in the long run (SS). As the labor market becomes more slack, hiring becomes easier; that is, recruiter productivity endogenously increases, making it easier for matches to occur given a fixed amount of recruitment effort. Under the carbon tax policies, recruitment initially increases in the clean sector as consumer substitution toward the clean good induces clean-sector firms to attempt to hire more workers. In the long run, recruitment declines in the clean sector, despite a larger workforce, because of the increased productivity of recruiters.

In the dirty sector, recruitment falls dramatically in the first period of the carbon tax policies as the firms reduce hiring below the replacement rate, causing the sector to shrink. But recruitment does not fall to zero, indicating that firms in the dirty sector decrease employment through slow hiring and natural attrition, rather than laying off workers.²⁴

Under any of the policy scenarios, there is almost no change in the probability of finding a job. Even in the lump-sum rebate carbon tax policy, where the change in job-finding probability is the largest, that change translates into an increase in average unemployment spells of just two days (compared with seven weeks in the baseline). Furthermore, the job-finding probability is almost flat across all policies over time, suggesting that workers looking for jobs immediately after the implementation of the policy spend the same amount of time looking for jobs as those looking for jobs years after the introduction of the policy.

Earnings in the dirty sector fall immediately after the introduction of environmental policy, with the drop most pronounced in the carbon tax policies. Workers in the clean sector also face decreases in earnings, but the initial drop is less severe. Over the longer term, earnings in both sectors converge to roughly the same level.

²⁴ This matches up with empirical results from Curtis (2012), who found that the drop in employment in regulated industries caused by the NO_x budget program occurred primarily through decreased hiring rates rather than increased job separations.

	Carbon tax lump-sum rebates	Carbon tax payroll tax cuts	Performance std.	
Increase in recruiter productivity				
Pd 0	4.00%	0.36%	0.51%	
SS	3.93%	0.35%	0.51%	
Recruiter ratio clean sector (% change)				
Pd 0	4.30%	11.02%	-1.47%	
SS	-3.62%	-0.34%	-0.49%	
Recruiter ratio dirty sector (% change)				
Pd 0	-53.68%	-48.40%	0.96%	
SS	-2.96%	-0.32%	-0.47%	
Increase in length of average				
unemployment spell (weeks)-SS	0.28	0.02	0.04	
Real after-tax earnings clean sector (% change)				
Pd 0	-0.71%	0.48%	-0.20%	
SS	-1.19%	-0.11%	-0.16%	
Real after-tax earnings dirty sector (% change)				
Pd 0	-3.72%	-2.57%	-0.12%	
SS	-1.21%	-0.14%	-0.20%	

Table 2. Labor Market Impacts

4.3. Welfare

Because we use a representative household framework, we cannot calculate the distribution of welfare effects across individual workers, but we can calculate aggregate welfare. The representative household utility function is given by

$$\overline{U}(c,h_c,h_d,n_c,n_d) = \log c - \psi n_c \frac{h_c^{1+\chi}}{1+\chi} - \psi n_d \frac{h_d^{1+\chi}}{1+\chi}$$

We use the equivalent variation (EV) measure of welfare changes. Using reference case prices, the EV is the amount of money that must be transferred to households such that \overline{U} (no policy plus transfers) = \overline{U} (policy). In our search model, consumers do not directly choose *n* or *h*. Employment is a function of recruiting effort of firms. Hours are bargained over with the firm. Therefore, in our EV calculations, we simply fix *n* and *h* at their no-policy levels. Note that because the utility function does not include external costs from pollution, this welfare measure effectively ignores any benefits of reduced emissions.

	Carbon tax lump-sum rebates	Carbon tax payroll tax cuts	Performance std.
EV	-\$8,692 b	-\$1,901 b	-\$2,746 b
EV (as % of wealth)	-0.13%	-0.03%	-0.04%
EV per ton reduced	-\$46.11	-\$10.44	-\$14.56

Table 3. Welfare

The welfare costs of the environmental policies are small, as shown in Table 3. The welfare cost of a \$20/ton carbon tax with lump-sum rebates is equivalent to a loss of 0.13 percent of wealth. When revenues are used to finance payroll tax cuts, the welfare loss is only 0.03 percent of wealth. As was the case with unemployment, the welfare costs of the performance standard are similar to the welfare costs of the carbon tax with payroll tax reductions.

These estimates might understate the cost of job dislocations. Davis and von Wachter (2011) show that workers who lose their jobs suffer not just a period of unemployment, but also a much longer-term drop in labor earnings. Our model includes the cost of that unemployment spell, but not any effect on earnings after a worker is re-employed. However, as noted earlier, we find that the employment shifts caused by environmental policy in our simulations happen entirely via natural attrition, rather than layoffs. Moreover, it is unclear how much (if any) of the earnings loss Davis and von Wachter found represents a social cost, rather than a purely distributional effect.²⁵

4.4. Stringency

Here we look at how the effects of the three policies change as we vary the stringency of the policies. Figure 4 displays the aggregate unemployment rate impacts of all three policies, as a function of the reduction in emissions relative to the no-policy steady state. At all levels of stringency, the rank order of the three policies remains the same. A carbon tax with lump-sum rebates has the highest unemployment rate increases, and a carbon tax with payroll tax reductions has the smallest increases. At very low levels of stringency, the unemployment rate

²⁵ For example, if the loss in labor earnings for laid-off workers represents rents that went to those workers before the layoffs and now flow to others in the economy, that is a loss for the laid-off workers but not for the economy as a whole.

impacts of a performance standard are very similar to those of a carbon tax with payroll tax reductions, though as the emissions reductions increase, the performance standard performs worse relative to the carbon tax with payroll tax cuts, because the disadvantage of relying almost entirely on reductions in emissions per unit of output (rather than reducing output from the polluting industry) grows as the policies become more stringent.





Figure 5 displays a similar result for the welfare cost (in terms of negative EV per ton reduced). The most cost-effective policy is the carbon tax with payroll tax cut policy. A performance standard is about as cost-effective at low levels of stringency but becomes much more costly as the emissions reductions increase. A carbon tax with lump-sum rebates is the least cost-effective policy for the entire range of stringency we consider (though we speculate that if we extended that range further, it could well be more cost-effective than the performance standard for very large reductions in emissions).²⁶

²⁶ These welfare-cost results and the way they change with stringency closely mirror results from the tax interaction literature (in models with full employment). See, for example, Goulder et al. (1999) or Goulder et al. (2016).



Figure 5: Welfare Costs by Stringency

5. Model Extensions

The aim of this section is to consider alternative extensions and/or specifications of the model to show how our results depend on some of these crucial modeling assumptions. We consider alternative specifications for wage bargaining and standard sensitivity analysis.

5.1. Wages

Thus far, we have allowed wages to be fully flexible and renegotiated each period. This assumption may be unrealistic for a variety of reasons, such as downward nominal wage rigidities, long-run contracts, or monopsony power in labor markets. Further, it is well known within the macroeconomic literature that search and matching models with fully flexible wages cannot replicate the historical volatility of unemployment in the United States.²⁷ As a result, we may be underestimating the impact of an emissions tax, because policy-induced declines in wages lead to smaller reductions in employment in the dirty sector (and more increases in employment in the clean sector) than would be the case if wages were not fully flexible.

To investigate the impact of flexible wages on our results, we consider staggered wage bargaining as in Gertler and Trigari (2009). To do this, we introduce a continuum of firms in

²⁷ For a complete analysis of the unemployment volatility puzzle, see Shimer (2010).

each sector. Firms are identical in each sector. Each period, a firm has a fixed probability $1 - \rho$ that it may renegotiate its wage contract.²⁸ The adjustment probability is independent over time, and although the length of contract is indeterminate, the average contract will last $1/(1-\rho)$ periods. We follow Gertler and Trigari and assume that all workers at a given firm receive the same wage. Given the Poisson adjustment and common-within-firm wage assumption, we can write the average wage in each sector as a function of an optimal target wage (which takes into consideration that the wage may be fixed for a lengthy period) and the average wage in the previous period.²⁹

Figure 6a displays the impact on aggregate unemployment of an emissions tax with our baseline flexible-wage model and two alternative staggered-wage models (with lump-sum rebates). The first staggered-wage model features an average length of a contract of three quarters (or nine months), and the second features an average length of a contract of nine quarters. Stickier wages cause a slower transition to the new steady state. In the long run, all wages can fully adjust, and thus the level of unemployment is the same as in the flexible model.³⁰ Figure 6b displays the impact of the performance standard policy (with labor tax adjustments) on unemployment. The transition looks very similar to that of the emissions tax with lump-sum rebates; the only difference is that the level of the change in unemployment is much smaller under the performance standard.³¹

²⁸ Unlike Gertler and Trigari, we allow for hours variation at each firm. We continue to assume that hours are set to maximize the joint surplus of a match. As a result, hours are equal across firms within a sector regardless of a given wage contract.

²⁹ An appendix with derivations of the staggered wage bargaining problem and the optimal wage dynamics is available from the authors on request.

³⁰ An alternative wage model would be to simply fix wages at pretax levels. This alternative model generates similar short-run and medium-run dynamics as the staggered-wage models but converges to a different steady state. We feel that long-run flexible wages are more realistic than permanently fixed wages.

³¹ If we instead raised payroll taxes in the performance standard, the sticky-wage model would generate a short-term spike in unemployment because of the higher payroll taxes.



Figure 6: Aggregate Unemployment and Sticky Wages

(a) Carbon Tax Lump-sum Rebates



(b) Performance Standard

The flexible- and sticky-wage models differ significantly in how the aggregate unemployment rate responds to changes in labor and payroll taxes. In particular, in the flexiblewage model, changes to labor and payroll taxes have the same unemployment impact because the wage can simply be negotiated to deliver the optimal take-home pay. When wages are fixed, even temporarily (as in the staggered-wage model), this is no longer true. A cut in labor income taxes will increase the take-home pay of the worker but will have no impact on the pretax wage paid by the employer. A reduction in employer payroll taxes, however, will decrease the net-of-tax payroll cost for the employer. This reduction in worker costs will increase the value of new hires to the firm, leading to a direct increase in both recruiting and jobs, relative to no payroll tax cut. These differences can be seen in Figure 7, which shows the response of aggregate unemployment to an emissions tax when revenues are recycled through either labor or employer payroll tax reductions in the three-quarter staggered-wage model.³² While both revenue-recycling measures lead to the same long-run unemployment rate very slightly above 7 percent (the same as in the flexible-wage model with revenue recycling), employer payroll tax reductions lead to a dramatically different transition as unemployment falls sharply by three months after the implementation of the emissions tax and payroll tax cut before slowly increasing over time.³³





 $^{^{32}}$ As in the flexible wage case, we consider a onetime permanent reduction in taxes such that the policy is fully revenue-neutral over the infinite horizon.

³³ This result may appear in part because our model allows employer recruitment effort to vary, but not workers' search effort. Thus making a successful job match more valuable to employers leads to more hiring (because employers respond by increasing recruiting effort), but making a successful match more valuable to employees causes no such effect.

5.2. Sensitivity Analysis

To illustrate how our results depend on specific parameter assumptions, Table 4 presents long-run unemployment and welfare loss per ton of emissions reductions (EV/ton) for our three policy scenarios as we vary key parameter values. In all of the simulations, we recalibrate disutility from work and unemployment benefits to maintain a steady-state unemployment rate of 7 percent in the absence of any environmental policy.

Table 4. Sensitivity Analysis						
	Carbon tax lump-sum rebates		Carbon tax	Carbon tax payroll tax cuts		nance std.
	\overline{u}	EV/ton	\overline{u}	EV/ton	\overline{u}	EV/ton
Benchmark	7.26	-\$46.11	7.02	-\$10.44	7.03	-\$14.56
Labor elasticity $1/\chi = 0.5$	7.32	-\$51.71	7.04	-\$12.01	7.05	-\$16.65
Labor elasticity $1/\chi = 2$	7.18	-\$43.11	7.01	-\$9.60	7.02	-\$13.51
Sector elasticity $\sigma^c = 0.5$	7.26	-\$48.69	7.02	-\$10.27	7.03	-\$13.28
Sector elasticity $\sigma^c = 1.5$	7.25	-\$39.96	7.03	-\$10.82	7.05	-\$18.28
Bargaining $\eta = 0.25$	7.07	-\$33.56	7.01	-\$8.71	7.01	-\$12.11
Bargaining $\eta = 0.75$	7.68	-\$74.91	7.15	-\$22.44	7.23	-\$32.30
Match elasticity $\gamma = 0.25$	7.38	-\$56.44	7.04	-\$12.34	7.06	-\$17.33
Match elasticity $\gamma = 0.75$	7.13	-\$35.26	7.01	-\$9.03	7.01	-\$12.50
Recruiter prod 50% lower	7.12	-\$35.56	7.01	-\$8.95	7.01	-\$12.47
Recruiter prod 50% higher	7.39	-\$56.09	7.04	-\$12.45	7.06	-\$17.44
50% lower taxes	7.26	-\$36.00	7.01	-\$8.74	7.02	-\$12.21
50% higher taxes	7.26	-\$55.19	7.05	-\$17.07	7.08	-\$23.69
Abatement cost 50% lower	7.24	-\$34.53	7.03	-\$10.22	7.04	-\$12.62
Abatement cost 50% higher	7.26	-\$54.00	7.02	-\$10.59	7.03	-\$16.17

We make three important observations based on the sensitivity analysis. First, emissions taxes with lump-sum rebates have a relatively small effect on aggregate unemployment in most cases (with the exception of high employer bargaining power). Second, across *all* simulations,

revenue recycling almost completely eliminates the impact of emissions pricing on the aggregate unemployment rate. Third, while the performance standard is always more costly than the carbon tax with payroll tax cuts, the unemployment rate of the performance standard is only slightly higher than the unemployment rate of the carbon tax with payroll tax cut policy.

For the simulations of the carbon tax with lump-sum revenue recycling, the effect of parameter values on long-run unemployment is consistent with our priors. The long-run unemployment impact is decreasing in the elasticity of labor supply. If workers are more willing to change their hours of work, more of the adjustment will be on the intensive margin, and thus less is needed on the extensive margin. Unemployment impacts of emissions taxes are largely independent of the elasticity of substitution between the two sectors in consumption; welfare cost per ton, on the other hand, decreases as the elasticity increases (making it easier to substitute from dirty to clean goods).

Bargaining power determines whether the firms or the workers bear a larger share of the incidence of the emissions tax. If employees have more bargaining power ($\eta < 0.5$)—in other words, they get a larger share of the surplus from a job match—then they also bear more of the burden of the tax. Thus employers' incentive to hire drops by less, and consequently the impact on unemployment is smaller. The opposite is true if employers have more bargaining power ($\eta > 0.5$). Changing the elasticity in the match functions changes the number of job matches that will be created for a given increase in recruiting effort. Decreasing the elasticity decreases the ability of the clean sector to hire workers who lost jobs in the dirty sector. As a result, the unemployment impact is decreasing in the match elasticity.

Increasing the steady-state recruiter productivity increases the long-run unemployment effect of any given environmental policy. The higher the recruiter productivity, the more any given change in the steady-state value of a job match will affect the steady-state level of unemployment (because with higher recruiter productivity, recruiting costs are lower, and thus it requires a larger change in labor-market tightness to offset any given change in the value of a match).

The higher the preexisting tax rate on labor, the larger is the welfare cost under each of the three policy scenarios (consistent with many results from full-employment models in the prior literature). The effect of a carbon tax with lump-sum rebates on unemployment is independent of the level of preexisting taxes, but the effect on unemployment is significantly

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larger under either the carbon tax with payroll tax cuts or the performance standard. All three policies cause the economy to contract slightly, which shrinks the tax base for the labor tax, thus reducing labor tax revenue (and the larger the preexisting labor tax, the bigger that drop is). Under the carbon tax with payroll tax cuts or the performance standard, the bigger that drop in revenue, the higher the payroll tax rate will be (because either the payroll tax cut is smaller or the increase is larger), so a larger preexisting labor tax implies a larger rise in unemployment. Under the carbon tax with lump-sum rebates, that simply means smaller lump-sum rebates, which leave unemployment largely unaffected.

As explained previously, a difference-in-differences estimator that assumes regulation does not affect clean sector employment would estimate the employment effect of an emissions tax as $\Delta_{dd} = \Delta n_d - \Delta n_c$ when the true effect on the dirty sector is Δn_d and the true total effect is $\Delta_u = \Delta n_d + \Delta n_c$. Notice that the dd estimate is equal to the true impact if and only if there is no employment change in the clean sector. That assumption is false in our simulations. Table 5 shows the values of Δ_{dd} and Δ_u in the benchmark scenario and the sensitivity scenarios, as well as the ratio of overestimation Δ_{dd} / Δ_u . This ratio is useful both as a measure of how misleading a dd estimator would be and also as a measure of how large the labor reallocation effect of the policy is relative to the net change in employment.

In all scenarios, there is significant reallocation of labor. Using emissions revenues to cut payroll taxes makes the drop in employment in the dirty sector smaller and the gain in the clean sector larger, decreasing the actual increase in unemployment but increasing the difference-in-differences estimate. As a result, a dd estimator would *dramatically* overestimate the employment impacts of an emissions tax coupled with revenue-neutral payroll tax cuts but would be off by much less (though still by a substantial margin) under the other two policy scenarios. Under the performance standard, looking only at dirty-sector employment gets the sign of the effect wrong (because employment rises in the dirty sector but falls by more in the clean sector). In that case, the dd estimator finds an even larger increase in employment, and thus is even further off.

The Δ_{dd} / Δ_u ratio is increasing in the labor supply elasticity, the elasticity of substitution between sector goods, and the match elasticity. The ratio is decreasing in bargaining power and recruiter productivity.

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	Carb	Carbon tax lump-		Carbon tax payroll		Performance std			
	Δ_u	Δ_{dd}	Ratio	Δ_u	Δ_{dd}	Ratio	Δ_u	Δ_{dd}	Ratio
Benchmark	0.255	0.788	3.09	0.023	0.930	40.98	0.033	-0.059	-1.77
Labor elasticity $1/\chi = 0.5$	0.317	0.752	2.37	0.035	0.925	26.79	0.050	-0.065	-1.29
Labor elasticity $1/\chi = 2$	0.180	0.831	4.62	0.014	0.932	67.54	0.020	-0.059	-2.87
Sector elasticity $\sigma^c = 0.5$	0.257	0.456	1.77	0.021	0.599	28.57	0.028	-0.071	-2.51
Sector elasticity $\sigma^c = 1.5$	0.251	1.772	7.07	0.028	1.910	68.69	0.049	0.022	0.45
Bargaining $\eta = 0.25$	0.071	0.888	12.59	0.005	0.928	196.48	0.007	-0.068	-9.89
Bargaining $\eta = 0.75$	0.682	0.532	0.78	0.149	0.857	5.77	0.227	-0.166	-0.73
Match elasticity $\gamma = 0.25$	0.375	0.715	1.91	0.040	0.919	22.99	0.059	-0.074	-1.26
Match elasticity $\gamma = 0.75$	0.130	0.865	6.63	0.010	0.938	94.89	0.014	-0.047	-3.28
Recruiter prod 50% lower	0.116	0.867	7.47	0.008	0.932	112.04	0.012	-0.059	-4.78
Recruiter prod 50% higher	0.385	0.711	1.85	0.042	0.920	21.88	0.062	-0.072	-1.16
50% lower taxes	0.255	0.788	3.09	0.014	0.935	68.03	0.020	-0.051	-2.53
50% higher taxes	0.255	0.788	3.09	0.053	0.911	17.20	0.077	-0.084	-1.09
Abatement cost 50% lower	0.244	0.733	3.00	0.030	0.863	28.29	0.039	-0.068	-1.76
Abatement cost 50% higher	0.260	0.812	3.12	0.044	0.959	49.53	0.031	-0.055	-1.77

Table 5. Overestimation of Aggregate Unemployment Effects

The Δ_{dd} / Δ_u ratio is weakly declining in the stringency of the policy. At a carbon tax of only \$5 per ton with the lump-sum rebates, the ratio is 3.37; at a tax of \$80 per ton with lump-sum rebates, the ratio is 2.37. Hence, the dd estimator overestimates the true employment impact regardless of policy stringency.

6. Conclusions

This paper used a simple two-sector general-equilibrium model with job search frictions to look at how environmental policies affect employment by sector and aggregate unemployment. It found that while imposing an environmental tax causes a substantial drop in employment in the polluting sector of the economy, it also indirectly causes an increase in employment in the nonpolluting sector. Thus the effect is primarily a shift of employment between sectors; the net effect on total employment (and unemployment) is small. And that net effect is even smaller if revenues from the tax are used to finance reductions in payroll taxes. An environmental performance standard results in a much smaller shift in employment between industries, though the net effect on employment is slightly larger than under a pollution tax with revenue recycled to cut payroll taxes.

These results suggest that overall effects on unemployment should not be a substantial factor in the evaluation of environmental policy. These effects are small and likely to be greatly outweighed by the other effects of the policy. However, because emissions taxes cause substantial sectoral shifts in employment, those effects could matter—and to the extent that the shifts are undesirable, that may provide a rationale for using performance standards (or other intensity-standard policies) rather than emissions taxes.

The results also suggest a need for caution in drawing lessons from partial-equilibrium empirical studies of the effects of environmental regulation. Such studies might accurately estimate the effects on the regulated industry, but not the effects on the rest of the economy, which our results suggest will be of roughly equal and opposite magnitude. Ignoring those effects and interpreting the estimated effect for the regulated industry as an estimate for the whole economy will thus lead to a drastic overestimate of the total effect (indeed, it could get the sign of that total effect wrong).

This problem is even more serious for difference-in-difference studies that use unregulated sectors as a control group: the rise in employment in those sectors will be mistakenly interpreted as a larger drop in employment in the regulated sectors.³⁴

However, this is not to say that those partial-equilibrium empirical studies are useless. Indeed, they could provide highly valuable information about the employment effects of regulation, as long as they are interpreted correctly (as the effect on the regulated sector or as the difference in effects between regulated and unregulated sectors). Taken together with a model like ours, such an estimate could be highly valuable.

³⁴ These papers often note this potential problem. For example, Greenstone (2002, p. 1212) notes, "Since it is likely that the regulation effects partially reflect some shifting of manufacturing activity within the United States, they probably *overstate* the national loss of activity due to the nonattainment designations. Moreover, the possibility of intra-country shifting means that the regulation effects are also likely to over-state losses in nonattainment counties." However, Greenstone then goes on to argue that this problem is likely small. And this caveat is almost always omitted when estimates from these papers are cited and used.

The general framework developed in this paper opens up many potential directions for future research. Perhaps the most interesting of these is whether environmental policy should vary across different stages of the business cycle, and if so, how. Suppose the government decides that a new environmental regulation is needed, but it reaches that decision during a major recession. Does it make sense to impose that new regulation immediately, or is it better to wait until after an economic recovery is well under way?

Another potential direction is to look at how unemployment changes the distributional effects of regulation. Even though unemployment has little influence on the aggregate welfare effects, it may have substantial distributional effects: workers in clean industries could be made substantially better off, while workers in dirty industries suffer. This would require a more disaggregated analysis, as well as adding occupation-specific skills (or switching costs to move between industries), but would otherwise fit well within the framework developed here.³⁵

The model could also readily be adapted to look at other types of policies. Trade policy is an obvious example; that is another setting in which employment effects are politically very important and in which CGE models are frequently used.

The transitions in the model are generally very quick, typically converging to a new steady state within less than two years. There are many reasons why real-world transitions might be slower, including the occupation-specific skills or switching costs just mentioned, as well as others such as capital adjustment costs or effects of work experience. Extending the model to include one or more of those factors would help add more realism to the transition dynamics.

Finally, it would be useful to empirically test some of the predictions of this model (or of an extended version of the model, such as the extension to occupation- or sector-specific skills). Clean identification would be a big challenge for testing many of the model's predictions, because a key aspect of the model is its general-equilibrium nature, and general-equilibrium effects are inherently difficult to identify empirically. Nonetheless, a careful empirical analysis could likely test at least some of the predictions from the model, and such a test would be quite valuable.

³⁵ Walker (2013) provides an interesting empirical analysis of what happens to workers in newly regulated firms, using a difference-in-difference-in-differences approach (conceptually very similar to Greenstone 2002). This approach has the same potential problem as other analyses using this approach: to the extent that general-equilibrium effects from regulation affect unregulated firms, this will bias the estimates of the effects of regulation. An extension of our model to include occupation- or sector-specific skills could be helpful by indicating how to interpret Walker's results (indicating the magnitude of the bias from general-equilibrium effects). At the same time, Walker's estimates could be useful as a way of calibrating that extended model or could serve as an empirical test of some of the predictions of such a model.

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